Caroline Vignet^{1†}, Tiziana Cappello², Qiuguo Fu¹, Kévin Lajoie³, Giuseppe De Marco²
 Christelle Clérandeau³, Hélène Mottaz¹, Maria Maisano², Juliane Hollender^{1,4}, Kristin
 Schirmer^{1,4*}, Jérôme Cachot^{3,*}

Imidacloprid induces adverse effects on fish early life stages
that are more severe in Japanese medaka (*Oryzias latipes*)
than in zebrafish (*Danio rerio*)

¹ Eawag, Swiss Federal Institute of Aquatic Science and Technology, 8600 Dübendorf,
8 Switzerland

9 ² University of Messina, Department of Chemical, Biological, Pharmaceutical and
10 Environmental Sciences, Messina 98166, Italy

³ Université de Bordeaux, Laboratoire EPOC, UMR CNRS 5805, 33615 Pessac Cedex,
 France

⁴ EPF Lausanne, School of Architecture, Civil and Environmental Engineering, 1015
 Lausanne, Switzerland and ETH Zurich, Institute of Biogeochemistry and Pollutant
 Dynamics, 8092 Zürich, Switzerland

- ¹⁶ [†] corresponding author actual adress: Institut national universitaire JF Champollion, Place de
- 17 Verdun, 81000 Albi, France. Email: caroline.vignet@inu-jfc.fr
- 18

This document is the accepted manuscript version of the following article:

Vignet, C., Cappello, T., Fu, Q., Lajoie, K., De Marco, G., Clérandeau, C., … Cachot, J. (2019). Imidacloprid induces adverse effects on fish early life stages that are more severe in Japanese medaka (Oryzias latipes) than in zebrafish (Danio rerio). Chemosphere. https://doi.org/10.1016/j.chemosphere.2019.03.002

This manuscript version is made available under the CC-BY-NC-ND 4.0 license http:// creativecommons.org/licenses/by-nc-nd/4.0/

19 ABSTRACT

Neonicotinoids are widely used insecticides that have frequently been found in freshwater 20 with concentrations ranging from ng to µg/L. It is known that these compounds impact non-21 target invertebrates, such as bees and gammaridae, in terms of toxicity and behavior, but 22 23 impacts and species differences on vertebrates such as fish are little explored. The aim of 24 this study was to investigate and compare the effects of one widely used neonicotinoid, imidacloprid, on development and behavior of two fish model species: Zebrafish (Danio rerio) 25 and Japanese medaka (Oryzias latipes). Fish were exposed for 5 (zebrafish) and 14 26 27 (medaka) days from 0.2 to 2000 µg/L imidacloprid by aqueous exposure. Survival, development, behavior and histological features were monitored and organism-internal 28 concentrations and biotransformation products measured. Imidacloprid caused sublethal 29 effects in both species but the effects were much stronger in medaka with deformities, 30 lesions and reduced growth being the most prominent impacts. Due to the overall longer time 31 of development, time-integrated exposure of medaka was about 2-fold higher compared to 32 zebrafish, potentially accounting for parts of the sensitivity differences. Our results underline 33 34 the importance of taking species sensitivity differences into account especially when 35 considering that medaka responded at imidacloprid concentrations that have been measured in the environment. 36

37 KEYWORDS: Cyprinids, toxicokinetics, species sensitivity, metabolome, embryo toxicity,
 38 imidacloprid.

39 HIGHLIGHTS:

- Imidacloprid impacts on fish at environmentally relevant concentrations.
- At the same developmental stage, medaka are more sensitive than zebrafish.
- Our study supports the importance of taking species sensitivity differences into account



45 **1. INTRODUCTION**

46 Neonicotinoids are one of the most produced pesticide families, even after the partial ban in Europe since 2013 (Van Dijk et al. 2013, Simon-Delso et al. 2014, Bonmatin et al. 2015, 47 Wood et al. 2017). They are low molecular weight and highly hydrophilic insecticidal 48 chemicals that are applied in agriculture in various ways, including foliar sprays and seed 49 50 treatments (Bonmatin et al. 2015). The intended mode of action of these molecules is to bind to nicotine acetylcholine receptors (nAChR) in nervous tissues in insects, causing 51 dysregulation of neurotransmission at cholinergic synapsis, which can lead to 52 overstimulation, tremors, paralysis and death (Sánchez-Bayo 2012, Simon-Delso et al. 53 54 2014). Yet, based on this mode of action, neonicotinoids could affect signal transmission and behavior of other organisms with a developed neuronal system (Sánchez-Bayo 2012). These 55 56 include animals living in aquatic environments (Tennekes 2011, Sánchez-Bayo 2012, Roessink et al. 2013, Sánchez-Bayo 2014). 57

Imidacloprid is one of frequently detected and well-studied neonicotinoids. It has been 58 detected up to several hundred µg/L after agricultural use but has most commonly been 59 found in the low ng/L range in continental water bodies (Moschet et al. 2014). Morissey et al. 60 61 (2015) reported up to 320 µg/L imidacloprid in drainage ditches in the Netherlands while Anderson et al.(2015) documented a peak concentration of 0.7 µg/L in a Canadian surface 62 water over a general background concentration of 0.04 to 0.05 µg/L. As demonstrated by 63 aquatic species sensitivity distribution on survival after a few days of exposure (SI Figure 64 65 S1), insects are the most vulnerable organism group, followed by crustaceans, while fish appear several orders of magnitude less sensitive to direct short-term exposure of 66 imidacloprid. Indirect effects on fish, such as a loss of the quantity and quality of crustaceans 67 serving as food (Hayasaka et al. 2012a, Gibbons et al. 2014, Chagnon et al. 2015), have 68 69 been proposed. However, direct sub-lethal effects on fish, especially during early developmental stages, have rarely been explored. Reduced locomotion was reported in 70 zebrafish larvae continuously exposed to imidacloprid from fertilization to five days (Crosby 71

et al. 2015). No impact was reported for zebrafish development when exposed to imidacloprid from fertilization to 48 hours (Tišler *et al.* 2009) and 96 hours (Scheil *et al.* 2009) of development, whereas growth of medaka adults and juveniles was reduced after long term exposure in mesocosms (Hayasaka *et al.* 2012b). Another study showed stress syndrome in medaka juvenile and increase parasite infestation after exposure to imidacloprid (Sanchez-Bayo *et al.* 2005). All these studies used exposure concentrations in the mg/L range, which is much higher than the concentrations found in the environment.

79 The aim of our study was to test if direct sub-lethal effects can be elicited by imidacloprid at concentrations that include environmentally realistic exposure levels during 80 81 critical stages of development, i.e. early life, in model fish species: Zebrafish (Danio rerio) and Japanese medaka (Oryzias latipes). While both species share common features such as 82 83 large broods, breeding all year and transparent eggs that develop outside the mother, a distinct difference is their time of development (SI Figure S2) (Kimmel et al. 1995, Furutani-84 85 Seiki et al. 2004, Iwamatsu 2004). While zebrafish hatch after 3 days post fertilization (dpf), medaka require an average of 9 dpf to emerge as free-swimming larvae. Thereafter, free 86 swimming larvae are completely established within 5 dpf in zebrafish where it takes 14 dpf in 87 medaka. We thus hypothesized that potential sub-lethal effects would be stronger in medaka 88 because of the longer developmental time and consequently greater time-integrated 89 exposure. To test this hypothesis, internal imidacloprid concentrations and physiological and 90 histological alterations were examined for both species at similar developmental stages. 91 Medaka indeed was more severely affected than zebrafish by imidacloprid exposure which 92 93 led us to explore the relative abundance of metabolites involved in energy metabolism and neurotransmission in this fish. 94

96

2. MATERIAL AND METHODS

97 2.1 Embryo collection

Zebrafish wild type (WT) were maintained and bred on site at the Eawag facility according to 98 the guidelines published by Nusslein-Volhard and Dahm, 2002 (Nüsslein-Volhard et al. 99 2002). Fish were raised at 28℃ in 14/10h light/dark cycle in reconstituted water (294.0 mg/L 100 101 CaCl₂ 2H₂O, 123.2 mg/L MgSO₄ 7H₂O, 64.74 mg/L NaHCO₃ and 5.7 mg/L KCl; prepared in 102 MilliQ water, pH 7.5) and fed twice daily with a combination of live food (Artemia nauplia) and 103 dry flakes (Tetramin, Switzerland). Adult fish were maintained in a large breeding tank (Aquatic habitat) with a special spawning system for collecting eggs. Eggs were collected 104 105 between 1 and 2 hours after the lights were turned on and fertilized eggs separated from unfertilized and placed in fresh medium in Petri dishes. Only spawns with more than 80% of 106 107 fertilized eggs were kept for the exposure. Medaka embryos were ordered from AMAGEN platform in Gif-sur-Yvette (France). They were transferred to the laboratory in hermetic boxes 108 109 and immediately used with exposure starting at 13 hpf. All procedures were in accordance with the animal protection guidelines. Experiments with zebrafish and medaka larvae were 110 approved by the Swiss Cantonal Veterinary Office (Number 119/2014) and by the French 111 ethic committee (Number A33-522-7), respectively. 112

113

114 2.2 Imidacloprid exposure

Based on concentrations reported from different water environments, the imidacloprid 115 exposure range was set from 0.2 to 2000 µg/L. Imidacloprid (PESTANAL[®], analytical 116 standard, Sigma-Aldrich 37894) was prepared as a 200 mg/L stock solution in 250 mL of 117 reconstituted water and aliquoted before being stored at -20°C. Serial dilutions with a factor 118 of 10 were prepared daily from this stock solution (see below). Three times twenty-five 119 120 fertilized embryos were placed in 3.5 cm Petri dishes with 3 mL of reconstituted water without 121 (control) or with imidacloprid. Exposure lasted for 5 (zebrafish) and 14 (medaka) days post fertilization (dpf). Water exchange was done every 24 hours to ensure stable aqueous 122 imidacloprid exposure concentrations (see chemical analysis below). Zebrafish were raised 123

124	at 28°C and medaka at 26°C in incubators (Economic Delux ECD01E model, Snijders
125	Scientific, Tilburg, NL for zebrafish and Memmert ICP 700 for medaka) in 14/10h light/dark
126	cycle in reconstituted water.
127	
128	2.3 Survival and development
129	Survival was monitored daily under the microscope and dead embryos or larvae were
130	removed. Fish were considered as dead when the heart did no longer beat. Percent survival
131	was calculated as compared to control for the last time point by using the ratio of survival
132	divided by the initial number of embryos.
133	
134	Hatch was likewise monitored daily. In controls, it is expected to occur around 3 days post
135	fertilization (dpf) for zebrafish (Kimmel et al. 1995) and 9 dpf for medaka (Iwamatsu 2004).
136	Hatchability was expressed as percent of control and calculated using the ratio of hatching
137	larvae divided by the initial number of embryos.
138	
139	After hatching (at 3 dpf for zebrafish and 9 dpf for medaka), 10 larvae per replicate and per
140	treatment were individually placed in 96 well plates. Microscope images of the whole body of
141	each larvae were taken for length measurement. Size was measured from mouth to end of
142	the tail. 10 larvae per biological replicate (3 replicates) were analyzed for a total of 30 fish per
143	treatment.
144	
145	Developmental anomalies were analyzed at the same time as size measurements under the
146	microscope following the protocol published by Le Bihanic (2013). Different types of
147	anomalies were recorded: heart, yolk-sac or bone oedema, tail problems (lordosis, kyphosis
148	or scoliosis), jaw or skull deformity, ocular lesions (missing eye, cyclopia and dystrophy),
149	heart curvature/position, hemorrhage and presence or absence of swollen swim bladder.

- 150 These results were expressed as percent compared to unexposed control.
- 151

152 2.4 Behavior

Behavioral experiments were performed at 5 dpf with zebrafish larvae (using Zebrabox from Viewpoint) and at 14 dpf with medaka larvae (using Daniovision from Noldus EthovisionXT11) and distance moved was video-tracked in both cases. First, fish were acclimatized in well plates in the dark for 2 hours at their optimal temperature (26°C for medaka and 28°C for zebrafish) before the test and then recorded for 3 periods. The first one was in the dark (light off-1; L.off-1), the second one in light (light on; L.on) and the third one in dark (light off-2; L.off-2).

160

For zebrafish, the procedure was as previously described ((Vignet *et al.* 2013, Vignet *et al.* 2015). The tested plate was transferred into the Zebrabox. Then 12 randommly selected zebrafish larvae per replicate per treatment were tested in 24 well plates in 5 min intervals. In light-off periods, zebrafish larvae normally present an increase of activity.

165

The procedure for medaka was as previously described in Granger Joly de Boissel *et al.* (2017). 10 larvae per replicate were randomly selected and placed individually in wells of a 48 well plate. After acclimatization, medaka larvae were video tracked in 10 min intervals. The last Light off (L.off-2) period represents the peak activity for medaka. According to Le Bihanic (2014) and Chiffre (2016), medaka activity is constant during the L.off-1 period and slightly increases during the L.on whereas for the last period, when the light is turned off again, larvae react with an increase of activity.

173

174 2.5 Histology

Eight samples for histological assessment were collected from control and each treatment at 5 dpf for zebrafish and at 14 dpf for medaka. The larvae were anesthetized with 0.01% MS 222 (Tricaine Methanesulfonate), and then immediately fixed in the Surgipath Decalcifier (Leica) for 24 h at 4°C. After dehydration in ethan ol, all specimens were embedded in paraffin (Bio-Optica, Italy) and sectioned at a thickness of 5 μm with a rotary automatic

microtome (Leica Microsystems, Wetzlar, Germany). Serial sections were stained with
hematoxylin and eosin (Bio-Optica, Isttaly), and examined with a motorized Zeiss Axio
Imager Z1 light microscope (Carl Zeiss AG, Werk Göttingen, Germany), equipped with an
AxioCam digital camera (Zeiss, Jena, Germany) for the acquisition of images. This protocol
was already published (Fasulo *et al.* 2010, Maisano *et al.* 2016, Maisano *et al.* 2017)

185

186 2.6 Liquid chromatography-high resolution mass spectrometry (LC-HRMS) analysis of
 187 imidacloprid and its biotransformation products in fish and exposure medium

Internal concentrations of imidacloprid and its biotransformation products were determined in 188 189 a separate set of experiments. A pool of 150 eggs was used in each of three independent experiments in which fish were raised up to 5 dpf for zebrafish and up to 14 dpf for medaka 190 191 in two Petri dishes with 20 mL of reconstituted water in each with 0 or 2000 µg/L of imidacloprid. Zebrafish were sampled after fertilization and at 3 dpf (hatching day) and 5 dpf 192 193 (larval stage and end of the exposure) while medaka were sampled after fertilization, at 3 dpf (same exposure days as zebrafish), 5 dpf (same exposure day as zebrafish), 9 dpf (hatching 194 day) and 14 dpf (larval stage and end of the exposure). The 150 embryos or larvae 195 (depending on the stage) per treatment were transferred into a cryotube and rinsed three 196 197 times with nanopure water. Water was removed with a pipet as much as possible and tubes were weighted and immediately flash-frozen in liquid nitrogen. 198

199

The sample preparation was based on the method by Rosch et al. (2017). In general, 100 µL 200 of 100 μ g/L imidacloprid d_4 , 500 μ L MeOH, and 300 mg of 1 mm zirconia/silica beads 201 (BioSpec Products, Inc., U.S.A.) were added to the frozen organisms, followed by 202 homogenization and extraction using a FastPrep bead beater (MP Biomedicals, Switzerland) 203 204 in two cycles (15 s, 6 m/s). The homogenized samples were centrifuged (6 min, 10 000 rpm, 205 20 ℃) and filtered through 0.45 µm regenerated cellulose filters (BGB Analytic AG, Switzerland). The supernatant was collected and the filters were washed with 400 µL MeOH. 206 The filtrate and the extract were eventually combined. 207

209 Imidacloprid concentration in embryo exposure medium was monitoring in the daily changed 210 solution and was sampled over the exposure to measure the effective imidacloprid 211 concentration (SI Table S1). All samples were stored at −20 °C until chemical analysis. The samples were analyzed by online solid phase extraction coupled to reversed phase liquid 212 chromatography high resolution tandem mass spectrometry (online-SPE-LC-HRMS/MS) (Q 213 214 Exactive, Thermo Fisher Scientific Inc.). Detection was done by full scan acquisition with a 215 resolution of 70000 (at m/z 200) in polarity switching mode (electrospray ionization) followed 216 by five (positive mode) and two (negative mode) data-dependent MS/MS scans with a resolution of 17000 (at m/z 200) with an isolation window of 1 m/z. Quantification was carried 217 out for imidacloprid and its metabolites using standards and imidacloprid-d4 as internal 218 219 standard.

220

221 2.7 NMR-based metabolomics analysis in medaka larvae

222 Endogenous polar metabolites from 14 dpf medaka larvae (n=3 pools of 20 fish each per group) were extracted using a "two-step" methanol/chloroform/water protocol, adequately 223 modified (Wu et al. 2008, Cappello et al. 2017b). In brief, medaka larvae were homogenized 224 225 in 8 mL/g of cold methanol and 2.5 mL/g of cold water by a TissueLyser LT bead mill 226 (Qiagen) with 0.5 mm glass beads, for 5 min at 50 vibrations/s, twice. Homogenates were transferred into glass vials, and 8 mL/g chloroform and 4 mL/g water were added. Samples 227 228 were vortexed for 30 s, and then incubated on ice for 10 min for phase separation. Following centrifugation at 2000 g for 5 min at 4 ℃, 200 µL of the upper methanol layer were 229 transferred into glass vials, dried in a centrifugal vacuum concentrator (Eppendorf 5301), and 230 231 kept at -80°C. Prior to Nuclear Magnetic Resonance (NMR) analysis, the dried polar extracts were resuspended in 600 µL of a 0.1 M sodium phosphate buffer (pH 7.0, 10% D₂O (Armar 232 AG, Döttingen, Switzerland)) containing 1 mM 2,2-dimethyl-2-silapentane-5-sulfonate (DSS) 233 (Sigma-Aldrich Co) used as internal standard, and then pipetted into a 5 mm NMR tube. 234

208

Extracts of medaka larvae were analyzed on a Varian-500 NMR spectrometer operating at a 236 spectral frequency of 499.74 MHz at 298 K. One-dimensional (1-D) ¹H NMR spectra were 237 238 obtained using a PRESAT pulse sequence to suppress the residual water resonance and 6,009 Hz spectral width with a 2.0 s relaxation delay. A total of 256 transients were collected 239 into 16,384 data points requiring a ca. 20 min acquisition time. All data sets were zero filled 240 to 32,768 data points and exponential line-broadenings of 0.5 Hz were applied before Fourier 241 242 transformation. All ¹H NMR spectra were manually phased, baseline-corrected, and calibrated (DSS at 0.0 ppm) using Chenomx NMR Suite (version 5.1; Chenomx Inc., 243 Edmonton, Canada) software. Peaks within the ¹H NMR spectra were assigned using the 244 Chenomx 500-MHz library and public databases. Chenomx NMR Suite was also used for 245 metabolite quantification (Cappello et al. 2017a, Cappello et al. 2017b, Maisano et al. 2017). 246

247

248 2.8 Statistical analysis

In order to test survival, hatch, length, deformity, and behavior, a General linear mixed model (GLM with Statistica software) was applied. When the results from GLM indicated a significant difference, a Newman-Keuls post hoc test was applied to compare groups. Significance levels were set at p < 0.05.

253

Statistical analyses for the metabolite data were conducted by the use of the GraphPad software (Prism 5.0, San Diego CA, USA). A one-way analysis of variance (ANOVA) was performed, followed by Dunnett's post-test, in order to determine the effects of single treatment groups compared to controls. Significance levels were set at p < 0.05.

258 259

```
260 3. RESULTS
```

261

Imidacloprid did not cause an impact on survival for any of the exposure conditions in eitherspecies of fish (data not shown). Moreover, overall hatching rate was unaffected for both

species despite a significant but transient alteration in hatching for some of the imidacloprid
concentrations at 7 dpf, and 8 dpf in medaka (SI Table S2). Yet, significant sub-lethal effects
were observed and these were more prominent for medaka as described below.

267

268 3.1 Impact on fish development and behavior

269

While no chemical-induced deformities and lesions were seen in zebrafish, the percentage of total anomalies in medaka reached 67% at the lowest tested imidacloprid concentration, 0.2 μ g/L, and >80% at 2 μ g/L and higher (Figure 1A). Lordosis/scoliosis, hemorrhage and jaw/skull deformity appeared in a concentration-dependent manner at $\geq 0.2 \mu$ g/L imidacloprid; oedema of the yolk and bones as well as tail deformities became visible at concentrations $\geq 20 \mu$ g/L (Figure 1B).

276

The stark difference in responding with sub-lethal effects between zebrafish and medaka 277 278 was also underlined in the histology. While no changes were found in the microscopic structure of the eves of zebrafish (Figure S3A, S3B), medaka larvae from 0.2 µg/L to the 279 highest concentration of imidacloprid exhibited a moderate disorganization of the retinal 280 281 pigment epithelium (Figure S3C, S3D). Additionally, a marked thickening of muscle fibers 282 was observed in zebrafish treated with 2000 µg/L of imidacloprid (Figure 2A, 2B) whereas in medaka an altered myomeric structure, as highlighted by heterogeneous alignment of the 283 fibers and presence of white spaces among them, was evident starting from the exposure of 284 \geq 2 µg/L of imidacloprid (Figure 2C, 2D). 285

286

Fish growth, measured as total length, was impacted by imidacloprid exposure in medaka but not in zebrafish. All imidacloprid exposed medaka larvae were about 5% smaller than the unexposed control group though this effect was not concentration-dependent (SI Table S3).

290

No impact by imidacloprid was found for both fish species for behavior. It pointed toward
hypoactivity though without a clear concentration-dependent pattern (Figure 3).

293

3.2 Toxicokinetics of imidacloprid and its biotransformation products in the developing fish

Imidacloprid was taken up by the developing fish. Yet, as demonstrated especially for medaka with its longer development phase, the chorion provided a significant barrier for uptake – only between 7 to 10% of the final organism-internal concentration were detected in the organisms at this early life stage. Imidacloprid was below the detection limit in control whereas concentrations in the exposed larvae drastically increased immediately after hatch, reaching 60 to 80% of the concentration measured in the larvae at the end of exposure (Table 1 column 1 and 2 and SI Table S4).

303

Apparent bioconcentration factors (BCFs) were calculated at the end of the exposures by dividing the concentration of imidacloprid in the organism by the concentration measured in the water recognizing that steady-state conditions may not have been reached. These BCFs amounted to 1.5 L/kg_{wet weight (ww)} and 1.2 L/kg_{ww} for medaka and zebrafish, respectively. In contrast in medaka just before hatch, the BCF was 0.1 L/kg_{ww}.

309

Of the total imidacloprid quantified at the respective end of the exposures, i.e. in 5 dpf 310 zebrafish and in 14 dpf medaka, about 15% were biotransformed as estimated based on the 311 presence of three biotransformation products (Table 1 and SI Table S4). The predominant 312 product in both species was hydroxyl-imidacloprid, which accounted for about 11% of 313 biotransformation. Urea-imidacloprid was found at less than 1% in both species. Finally, 314 315 olefin-imidacloprid represented 1% in zebrafish and about 3% in medaka. Desnitro-316 imidacloprid, which has been described in bees (Suchail et al. 2001), was not detected in 317 either species.

319 3.3 Metabolic responses in medaka

320

To better understand the responses of medaka to imidacloprid exposure, we monitored the levels of endogenous metabolites in medaka larvae. Compared with larvae from controls (see SI Figure S4 for a representative 1-D ¹H NMR spectrum of 14 dpf medaka larvae), the exposure to imidacloprid resulted in significant changes (p < 0.05) in metabolites which can be assigned to either energy metabolism (namely glucose, pyruvate, succinate, ATP/ADP, and lactate) or to cholinergic (choline and acetylcholine) and to adrenergic (tyrosine and phenylalanine) neurotransmission (Table 2 and SI Figure S4).

328

329 4. DISCUSSION

The aim of our study was two-fold: (1) to explore if sub-lethal effects, ranging from impacts on development to behavior, can arise from imidacloprid early life stage exposure of two model fish – zebrafish and medaka – embracing concentrations close to environmental levels; and (2) to test if species differences arise.

Imidacloprid caused sublethal effects in both species but the effects were much more severe 334 335 in medaka. The most prominent impact was the induction of deformities and lesions. The mechanisms leading to such effects can be manifold and will need further investigations to 336 be precisely understood. However, several lines of evidence point toward an involvement of 337 338 nicotinic acetylcholine receptors (nAChRs), i.e. the specific target of imidacloprid in insects, keeping in mind that, while in insects nAChRs are restricted to the central nervous system, 339 340 these same receptors are also present at neuromuscular junctions in vertebrates (Millar et al. 2009). 341

342

The first line of evidence for an involvement of nAChRs is that the anomalies herein observed in the muscle structure of medaka larvae, with the heterogeneous alignment of the fibers, could be explained by a dysregulation of muscle contractions due to interference by

imidacloprid with the fish neuromuscular nAChRs. The exposure to imidacloprid can provoke 346 muscle contraction with consequent release of lactate, as evidenced by an increase in 347 348 lactate levels at the lowest exposure concentrations. Increased levels of acetylcholine comprise the second line of evidence of the interference of imidacloprid with the fish 349 nAChRs. Indeed, increased acetylcholine levels recorded in all the exposure groups are in 350 agreement with the action and target-site selectivity of imidacloprid that saturates nAChRs 351 352 (Matsuda et al. 2009), leading to an increase of free acetylcholine. Similar effects can be 353 elicited in zebrafish, as demonstrated by Tufi et al (2016), but only at concentrations at least one order of magnitude higher than the highest concentration tested here. Along the lines of 354 these sensitivity differences, the marked thickening of muscle fibers in zebrafish larvae at the 355 highest concentration of imidacloprid in our study was 2000-fold higher than the lowest 356 concentration at which this effect was observable in medaka. We therefore conclude that the 357 neuromuscular nAChRs might be an important target of imidacloprid especially during early 358 developmental stages in both species. The difference between medaka and zebrafish could 359 360 be due to a greater affinity of the medaka nAChRs to imidacloprid or an overall higher activation of nAChRs due to greater time-integrated imidacloprid exposure levels in the 361 developing medaka. Moreover, besides the alterations in the cholinergic system, 362 disturbances in the adrenergic system were also detected herein in medaka, with increased 363 level of phenylalanine, a precursor of tyrosine, which was, however, not followed by an 364 increase in the level of tyrosine itself. Similar data were observed in zebrafish larvae after 365 366 exposure to imidacloprid, which induced increased levels of phenylalanine but no change in 367 the levels of tyrosine (Tufi et al. 2016).

368

Zebrafish and medaka are similar in many traits, such as size, optimal temperature range
and being oviparous (Furutani-Seiki *et al.* 2004). However, the developmental time from
fertilization to free swimming larvae is about three times longer in medaka than in zebrafish
(14 d for medaka, 5 d for zebrafish) (Kimmel *et al.* 1995, Furutani-Seiki *et al.* 2004, Iwamatsu
2004)

At time of hatching (9 d for medaka, 3 d for zebrafish), medaka larvae contained about twice 374 the amount of imidacloprid per g of tissue compared to zebrafish larvae. If passive uptake 375 376 into the embryo is assumed, the higher accumulation in medaka larvae could be explained 377 by the longer exposure time, assuming that steady-state-concentrations have not been reached. Yet another factor influencing accumulation could be the difference the composition 378 379 of the chorion between medaka and zebrafish. While both zebrafish and medaka have a transparent chorion with 3 layers (Bonsignorio et al. 1996), there are differences in certain 380 381 traits. The chorion of zebrafish is soft compared to the chorion of medaka and is as well smooth compared to the hair structures on the medaka chorion surface. The relative 382 hardness of the medaka chorion could be due to a higher content of proline (Bonsignorio et 383 al. 1996), which is known to contributes to the structural stability of proteins like collagen 384 (Jaeken 2012). Such features, i.e. biochemical compositions and surface structures like hair, 385 could contribute to the medaka chorion being more prone to chemical uptake than the 386 chorion of zebrafish. 387

388 After hatch, imidacloprid concentrations rose in both species of fish though overall apparent BCFs indicate that imidacloprid does not accumulate strongly in both species (1.5 L/kgwet weight 389 (ww) and 1.2 L/kgww for medaka and zebrafish, respectively). This is in accordance with BCFs 390 391 for imidacloprid reported for other species of fish (australoheros facetus (lturburu et al. 2016) 392 (1.4L L/kgww)) or amphibians (Van Meter et al. 2016) (0.2 to 0.7L L/kgww) whereas in gammarids, the BCF was higher (7.35 L/kg_{ww}) (Ashauer et al. 2010, Ashauer et al. 2012). At 393 the end of the exposure, medaka showed a 1.3-fold higher accumulation compared to 394 zebrafish, which contrasts with the orders of magnitude difference in developmental 395 sensitivity of the medaka compared to zebrafish. Thus, while longer exposure time, paired 396 397 with greater internal exposure, are conceivable contributors to the comparatively high sensitivity of medaka, other factors appear to contribute to the species sensitivity differences. 398 One such factor could be biotransformation though the only notable difference was in the 399 formation of olefin-imidacloprid. Interestingly, olefin-imidacloprid was found to be twice as 400 lethally toxic than the parent compound in bees after 48 hours of exposure (Suchail et al. 401

2001). It was also shown to be potentially more toxic in mice (Lee Chao *et al.* 1997,
Tomizawa *et al.* 1999). Thus, exploration of sensitivity toward olefin-imidacloprid is one future
avenue to shed light on the species differences observed.

405

The comparatively high sensitivity of medaka toward imidacloprid during early development 406 407 underlines the importance of taking species differences for environmental risk assessment into account. As demonstrated in the species sensitivity distribution (SI Figure S1), only few 408 species of fish have thus far been explored for their sensitivity to imidacloprid with a focus on 409 acute exposure. Though the exact mechanisms of the high sensitivity of the medaka during 410 411 early life stages still need to be further explored, this fish appears about three orders of magnitude more sensitive to imidacloprid than the zebrafish. The most important impacts 412 413 measured herein with regard to ecological relevance are the developmental toxicity and the reduced growth of medaka. Both these types of impact can conceivably be linked. For 414 415 example, the alteration of muscle fibers can result in reduced locomotion, which in turn can result in reduced ability to catch food. This thought is supported by the observed tendency 416 toward hypoactive behavior and disturbances in the neurotransmission pathways, both in the 417 cholinergic and adrenergic systems. Similarly, alterations to the structure of the eyes may 418 419 obstruct perception of predator or prey. All these impacts therefore can severely hamper the fitness of the fish in their natural environment. In the future, research priorities to further 420 explore the species sensitivity differences could be to *i*) explore the barrier function of the 421 chorion on chemical uptake, ii) mechanisms of neuromuscular nAChRs and iii) the toxicity of 422 423 olefin-imidacloprid in both species. In this regard it is important to note that 0.2 µg/L of imidacloprid, i.e. the lowest concentration at which strong effects were already seen in 424 425 medaka, is in the range of concentrations (μ g/L) reported in some environments like rivers, groundwaters, streams and estuaries (Anderson et al. 2015, Morrissey et al. 2015). 426

427

428

	ACCEPTED MANUSCRIPT
429	
430	
431	5. Acknowledgement
432	Caroline Vignet was supported by Eawag funding. Qiuguo Fu was supported by the Swiss
433	National Science Foundation, grant No 205320_165935. We would like to thanks for Pascal
434	Reichlin and other members of the Eawag Department of Environmental Toxicology for fish
435	maintenance and embryo production.
436	
437	6. Declarations of interest
438	None.
439	
440	



Figure 1. Deformities in medaka at hatch (9 dph). A. Percentage of total developmental anomalies. Stars indicate significant differences to control fish (0µg/L) for p ≤ 0.05 as determined by GLM. B. Anomalies are ranked into three categories. Group 1: non chemical-specific anomalies (plain grey to dark); they include lack of swim bladder inflation, ocular lesion (missing eye, cyclopia and dystrophy), heart oedema and heart position/curvature. Group 2: concentration dependent anomalies (dashed line patterns); they include lordosis/scoliosis, hemorrhage and jaw/skull deformity. Group 3: anomalies that became visible only at concentrations $\ge 20 \ \mu g/L$ (dotted patterns); they include oedema of the yolk and bones as well as tail deformity. These experiments was done 3 times with 10 embryos per treatment each time.

Control

Imidacloprid

В

D



Zebrafish

Medaka

- 461 Figure 2. Histological sections of muscles of 5 dpf zebrafish (A-B) and 14 dpf medaka (C-D), stained with Hematoxylin and Eosin. A regular
- 462 microscopic structure of the muscles of zebrafish from control group (A) and marked thickening of muscle fibers (double ended arrow) in zebrafish
- 463 treated with 2000 µg/L of imidacloprid (B). Muscles of medaka from control (C) and a representative image showing an altered myomeric structure
- with heterogeneous alignment of the fibers (arrows), and presence of white spaces among them, found from 2 µg/L to the highest concentration of
- 465 imidacloprid group (D). nc, nothocord; m, muscle. Scale bars, 20 μm.



Figure 3. Behavioral response in zebrafish after 5 days of exposure (A, B) and in medaka after 14 days of exposure (C, D) to different
concentrations of imidacloprid. (A). Distance moved measured every 30 sec. (B). Distance moved measured every 5 minutes. (C). Distance moved
measured every minute. (D). Distance moved measured every 10 minutes. Zf-Loff1: zebrafish Light off1 (5 min); Zf Lon: zebrafish Light On (5 min);
Zf-Loff2: zebrafish light off 2 (5 min); m-Loff1: medaka Light off1 (10 min); m-Lon: medaka Light On (10 min); m-Loff2: medaka light off 2 (10 min).

- **Table 1**: Whole body internal concentration of imidacloprid and biotransformation products in ng/g w.t after imidacloprid exposure (n.d= not
- 473 detected; n.q= detected but under limit of quantification (see Table S4).

		Imidacloprid	Hydroxyl- imidacloprid	Desnitro- imidacloprid	Olefin- imidacloprid	Urea- imidacloprid
	Unexposed larvae (3, 5 dpf)	n.d	n.d	n.d	n.d	n.d
zobrofich	Larvae at hatching day (3 dpf)	1267 ± 58	64 ± 4	n.q	9 ± 1	8 ± 0.5
Zepransn	Larvae at the end of experiment (5 dpf)	2067 ± 153	263 ± 31	n.q	22 ± 3	11 ± 1
	Unexposed embryo (3, 5 dpf) and larvae (9,14 dpf)	n.d	n.d	n.d	n.d	n.d
	3 dpf (embryo)	180 ± 35	n.d	n.d	n.d	n.d
medaka	5 dpf (embryo)	273 ± 12	n.d	n.d	n.d	n.q
	Larvae at hatching day (9 dpf)	2133 ± 115	160 ± 20	n.d	24 ± 3	11 ± 1
	Larvae at the end of experiment (14 dpf)	2667 ± 252	390 ± 36	n.d	76 ± 8	12 ± 1

	0.2 μg/L	2 μg/L	20 μg/L	200 μg/L	2000 μg/L					
Metabolites relative to energy metabolism										
Glucose	↓ 10%	↓ 13%	↓ 24%	₩ 38%	↓ 31%					
Pyruvate	♠ 64%*	↑ 70%*	↑ 70%*	▲ 39%	个 43%					
Succinate	♠ 62%*	↑ 75%*	↑ 58%*	个 14%	个 4%					
ATP/ADP	↑ 30%	1 42%*	1 40%*	↓ 14%	no change					
Lactate	♠ 52%*	1 47%*	♠ 50%*	↓ 46%*	↓ 21%					
Metabolites relative to cholinergic neurotransmission										
Choline	1 24%	1 40%*	↑ 16%	↓ 23%	↓ 12%					
Acetylcholine	♠ 64%*	↑ 44%*	↑ 34%	个 25%	个 17%					
Met	tabolites relativ	ve to adrene	rgic neurotra	nsmission						
Tyrosine	♠ 66%*	1 29%	↑ 16%	↓ 10%	no change					
Phenylalanine	↑ 78%*	♠ 70%*	↑ 104%*	♠ 68%*	个 37%					

Table 2. Percent changes in concentrations of metabolites between imidacloprid-exposed and control medaka (Dunnett's test; **p* < 0.05).

480

481

Anderson, J. C., C. Dubetz and V. P. Palace (2015). "Neonicotinoids in the Canadian aquatic environment: a literature review on current use products with a focus on fate, exposure, and biological effects." <u>Sci Total Environ</u> **505**: 409-422.

Ashauer, R., I. Caravatti, A. Hintermeister and B. I. Escher (2010). "Bioaccumulation kinetics of organic
xenobiotic pollutants in the freshwater invertebrate *Gammarus pulex* modeled with prediction
intervals." <u>Environ Toxicol Chem</u> 29(7): 1625-1636.

Ashauer, R., A. Hintermeister, I. O'Connor, M. Elumelu, J. Hollender and B. I. Escher (2012).
"Significance of Xenobiotic Metabolism for Bioaccumulation Kinetics of Organic Chemicals in *Gammarus pulex.*" <u>Environmental Science & Technology</u> **46**(6): 3498-3508.

Bonmatin, J. M., C. Giorio, V. Girolami, D. Goulson, D. P. Kreutzweiser, C. Krupke, M. Liess, E. Long, M.
Marzaro, E. A. Mitchell, D. A. Noome, N. Simon-Delso and A. Tapparo (2015). "Environmental fate
and exposure; neonicotinoids and fipronil." <u>Environ Sci Pollut Res Int</u> 22(1): 35-67.

494 Bonsignorio, D., L. Perego, L. Del Giacco and F. Cotelli (1996). "Structure and macromolecular 495 composition of the zebrafish egg chorion." <u>Zygote</u> **4**(2): 101-108.

Cappello, T., M. Maisano, A. Mauceri and S. Fasulo (2017a). "1H NMR-based metabolomics
investigation on the effects of petrochemical contamination in posterior adductor muscles of caged
mussel *Mytilus galloprovincialis*." <u>Ecotoxicol Environ Saf</u> 142: 417-422.

Cappello, T., V. Vitale, S. Oliva, V. Villari, A. Mauceri, S. Fasulo and M. Maisano (2017b). "Alteration of
 neurotransmission and skeletogenesis in sea urchin *Arbacia lixula* embryos exposed to copper oxide
 nanoparticles." <u>Comp Biochem Physiol C Toxicol Pharmacol</u>.

502 Chagnon, M., D. Kreutzweiser, E. D. Mitchell, C. Morrissey, D. Noome and J. Van der Sluijs (2015).
503 "Risks of large-scale use of systemic insecticides to ecosystem functioning and services."
504 <u>Environmental Science and Pollution Research</u> 22(1): 119-134.

Chiffre, A., C. Clerandeau, C. Dwoinikoff, F. Le Bihanic, H. Budzinski, F. Geret and J. Cachot (2016).
"Psychotropic drugs in mixture alter swimming behaviour of Japanese medaka (*Oryzias latipes*) larvae
above environmental concentrations." <u>Environ Sci Pollut Res Int</u> 23(6): 4964-4977.

508 Crosby, E. B., J. M. Bailey, A. N. Oliveri and E. D. Levin (2015). "Neurobehavioral impairments caused 509 by developmental imidacloprid exposure in zebrafish." <u>Neurotoxicology and Teratology</u>(0).

Fasulo, S., A. Mauceri, M. Maisano, A. Giannetto, V. Parrino, F. Gennuso and A. D'Agata (2010).
"Immunohistochemical and molecular biomarkers in *Coris julis* exposed to environmental
contaminants." <u>Ecotoxicology and Environmental Safety</u> **73**(5): 873-882.

Furutani-Seiki, M. and J. Wittbrodt (2004). "Medaka and zebrafish, an evolutionary twin study." <u>Mech</u>
 <u>Dev</u> 121(7-8): 629-637.

515 Gibbons, D., C. Morrissey and P. Mineau (2014). "A review of the direct and indirect effects of 516 neonicotinoids and fipronil on vertebrate wildlife." <u>Environmental Science and Pollution Research</u>: 1-517 16.

518 Granger Joly de Boissel, P., P. Gonzalez, A. Bulete, G. Daffe, C. Clerandeau, E. Vulliet and J. Cachot 519 (2017). "An innovative and integrative assay for toxicity testing using individual fish embryos. 520 Application to oxazepam." <u>Chemosphere</u> **181**: 468-477.

521 Hayasaka, D., T. Korenaga, F. Sanchez-Bayo and K. Goka (2012a). "Differences in ecological impacts of

systemic insecticides with different physicochemical properties on biocenosis of experimental paddy
 fields." <u>Ecotoxicology</u> 21(1): 191-201.

Hayasaka, D., T. Korenaga, K. Suzuki, F. Saito, F. Sanchez-Bayo and K. Goka (2012b). "Cumulative
ecological impacts of two successive annual treatments of imidacloprid and fipronil on aquatic
communities of paddy mesocosms." <u>Ecotoxicol Environ Saf</u> 80: 355-362.

527 Iturburu, F. G., M. Zomisch, A. M. Panzeri, A. C. Crupkin, V. Contardo-Jara, S. Pflugmacher and M. L.

528 Menone (2016). "Uptake, distribution in different tissues, and genotoxicity of imidacloprid in the 529 freshwater fish Australaberos facetus." Environ Toxicol Chem

529 freshwater fish *Australoheros facetus*." <u>Environ Toxicol Chem</u>.

- Iwamatsu, T. (2004). "Stages of normal development in the medaka *Oryzias latipes*." <u>Mechanisms of</u>
 <u>Development</u> 121(7–8): 605-618.
- 532 Jaeken, J. (2012). Disorders of Proline and Serine Metabolism. <u>Inborn Metabolic Diseases: Diagnosis</u>
- and Treatment. J.-M. Saudubray, G. van den Berghe and J. H. Walter. Berlin, Heidelberg, Springer
 Berlin Heidelberg: 357-362.
- 535 Kimmel, C. B., W. W. Ballard, S. R. Kimmel, B. Ullmann and T. F. Schilling (1995). "Stages of embryonic 536 development of the zebrafish." <u>Developmental Dynamics</u> **203**(3): 253-310.
- 537 Le Bihanic, F. (2013). Effects of polycyclic aromatic hydrocarbons on early life stages of model fish :
- 538 <u>development of bioessay and comparative study of mixtures</u>, Université Sciences et Technologies 539 Bordeaux I.
- Le Bihanic, F., C. Clérandeau, K. Le Menach, B. Morin, H. Budzinski, X. Cousin and J. Cachot (2014). "Developmental toxicity of PAH mixtures in fish early life stages. Part II: adverse effects in Japanese
- 542 medaka." <u>Environmental science pollution research</u>.
- Lee Chao, S. and J. E. Casida (1997). "Interaction of Imidacloprid Metabolites and Analogs with the
 Nicotinic Acetylcholine Receptor of Mouse Brain in Relation to Toxicity." <u>Pesticide Biochemistry and</u>
 <u>Physiology</u> 58(1): 77-88.
- 546 Maisano, M., T. Cappello, A. Natalotto, V. Vitale, V. Parrino, A. Giannetto, S. Oliva, G. Mancini, S.
- 547 Cappello, A. Mauceri and S. Fasulo (2017). "Effects of petrochemical contamination on caged marine
- mussels using a multi-biomarker approach: Histological changes, neurotoxicity and hypoxic stress."
 Marine Environmental Research 128: 114-123.
- 550 Maisano, M., T. Cappello, S. Oliva, A. Natalotto, A. Giannetto, V. Parrino, P. Battaglia, T. Romeo, A.
- Salvo, N. Spanò and A. Mauceri (2016). "PCB and OCP accumulation and evidence of hepatic
 alteration in the Atlantic bluefin tuna, *T. thynnus*, from the Mediterranean Sea." <u>Marine</u>
 <u>Environmental Research</u> 121: 40-48.
- 554 Matsuda, K., S. Kanaoka, M. Akamatsu and D. B. Sattelle (2009). "Diverse actions and target-site 555 selectivity of neonicotinoids: structural insights." <u>Mol Pharmacol</u> **76**(1): 1-10.
- 556 Millar, N. S. and C. Gotti (2009). "Diversity of vertebrate nicotinic acetylcholine receptors."
 557 <u>Neuropharmacology</u> 56(1): 237-246.
- 558 Morrissey, C. A., P. Mineau, J. H. Devries, F. Sanchez-Bayo, M. Liess, M. C. Cavallaro and K. Liber 559 (2015). "Neonicotinoid contamination of global surface waters and associated risk to aquatic 560 invertebrates: a review." <u>Environ Int</u> **74**: 291-303.
- 561 Moschet, C., I. Wittmer, J. Simovic, M. Junghans, A. Piazzoli, H. Singer, C. Stamm, C. Leu and J. 562 Hollender (2014). "How a Complete Pesticide Screening Changes the Assessment of Surface Water 563 Quality." Environmental Science & Technology **48**(10): 5423-5432.
- 564 Nüsslein-Volhard, C. and R. Dahm (2002). <u>Zebrafish, a technical approach</u>. Tübingen, Oxford 565 University Press.
- Roessink, I., L. B. Merga, H. J. Zweers and P. J. Van den Brink (2013). "The neonicotinoid imidacloprid
 shows high chronic toxicity to mayfly nymphs." <u>Environ Toxicol Chem</u> **32**(5): 1096-1100.
- Rosch, A., M. Gottardi, C. Vignet, N. Cedergreen and J. Hollender (2017). "Mechanistic Understanding
 of the Synergistic Potential of Azole Fungicides in the Aquatic Invertebrate *Gammarus pulex*." <u>Environ</u>
 <u>Sci Technol</u>.
- 571 Sánchez-Bayo, F. (2012). "Insecticides Mode of Action in Relation to Their Toxicity to Non-Target 572 Organisms." J Environment Analytic Toxicol **\$4-002**.
- 573 Sánchez-Bayo, F. (2014). "The trouble with neonicotinoids." <u>Science</u> **346**(6211): 806-807.
- 574 Sanchez-Bayo, F. and K. Goka (2005). "Unexpected effects of zinc pyrithione and imidacloprid on 575 Japanese medaka fish (*Oryzias latipes*)." <u>Aquat Toxicol</u> **74**(4): 285-293.
- 576 Scheil, V. and H. R. Kohler (2009). "Influence of nickel chloride, chlorpyrifos, and imidacloprid in
- 577 combination with different temperatures on the embryogenesis of the zebrafish *Danio rerio*." <u>Arch</u>
 578 <u>Environ Contam Toxicol</u> 56(2): 238-243.
- 579 Simon-Delso, N., V. Amaral-Rogers, L. P. Belzunces, J. M. Bonmatin, M. Chagnon, C. Downs, L. Furlan,
- 580 D. W. Gibbons, C. Giorio, V. Girolami, D. Goulson, D. P. Kreutzweiser, C. H. Krupke, M. Liess, E. Long,
- 581 M. McField, P. Mineau, E. A. Mitchell, C. A. Morrissey, D. A. Noome, L. Pisa, J. Settele, J. D. Stark, A.

- Tapparo, H. Van Dyck, J. Van Praagh, J. P. Van der Sluijs, P. R. Whitehorn and M. Wiemers (2014).
 "Systemic insecticides (neonicotinoids and fipronil): trends, uses, mode of action and metabolites."
 Environ Sci Pollut Res Int.
- 585 Suchail, S., D. Guez and L. P. Belzunces (2001). "Discrepancy between acute and chronic toxicity 586 induced by imidacloprid and its metabolites in *Apis mellifera*." <u>Environ Toxicol Chem</u> **20**(11): 2482-587 2486.
- 588 Tennekes, H. A. (2011). "The significance of the Druckrey–Küpfmüller equation for risk assessment—
- 589 The toxicity of neonicotinoid insecticides to arthropods is reinforced by exposure time: Responding
- to a Letter to the Editor by Drs. C. Maus and R. Nauen of Bayer CropScience AG." <u>Toxicology</u> 280(3):
 173-175.
- 592 Tišler, T., A. Jemec, B. Mozetič and P. Trebše (2009). "Hazard identification of imidacloprid to aquatic 593 environment." <u>Chemosphere</u> **76**(7): 907-914.
- 594 Tomizawa, M. and J. E. Casida (1999). "Minor structural changes in nicotinoid insecticides confer 595 differential subtype selectivity for mammalian nicotinic acetylcholine receptors." <u>Br J Pharmacol</u> 596 **127**(1): 115-122.
- 597 Tufi, S., P. Leonards, M. Lamoree, J. de Boer, J. Legler and J. Legradi (2016). "Changes in
- 598Neurotransmitter Profiles during Early Zebrafish (Danio rerio) Development and after Pesticide599Exposure." Environ Sci Technol **50**(6): 3222-3230.
- Van Dijk, T. C., M. A. Van Staalduinen and J. P. Van der Sluijs (2013). "Macro-Invertebrate Decline in
 Surface Water Polluted with Imidacloprid." <u>PLoS ONE</u> 8(5): e62374.
- Van Meter, R. J., D. A. Glinski, W. M. Henderson and S. T. Purucker (2016). "Soil organic matter
 content effects on dermal pesticide bioconcentration in American toads (*Bufo americanus*)."
 Environmental Toxicology and Chemistry **35**(11): 2734-2741.
- Vignet, C., M. Bégout, S. Péan, L. Lyphout, D. Leguay and X. Cousin (2013). "Systematic screening of
 behavioral responses in two zebrafish strains." <u>Zebrafish.</u> Sep;10(3):365-75.
- 607 Vignet, C., L. Joassard, L. Lyphout, T. Guionnet, M. Goubeau, K. Le Menach, F. Brion, O. Kah, B. C.
- 608 Chung, H. Budzinski, M. L. Begout and X. Cousin (2015). "Exposures of zebrafish through diet to three
 609 environmentally relevant mixtures of PAHs produce behavioral disruptions in unexposed F1 and F2
 610 descendant." <u>Environ Sci Pollut Res Int</u> 21: 16371-16383.
- 611 Wood, T. J. and D. Goulson (2017). "The environmental risks of neonicotinoid pesticides: a review of 612 the evidence post 2013." <u>Environ Sci Pollut Res Int</u>.
- 613 Wu, H., A. D. Southam, A. Hines and M. R. Viant (2008). "High-throughput tissue extraction protocol 614 for NMR- and MS-based metabolomics." <u>Anal Biochem</u> **372**(2): 204-212.
- 615

HIGHLIGHTS:

- Imidacloprid impacts on fish at environmentally relevant concentrations.
- At the same developmental stage, medaka are more sensitive than zebrafish.
- Our study supports the importance of taking species sensitivity differences into account

