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Proteome changes in muscles, ganglia, and gills
in *Corbicula fluminea* clams exposed to crude oil:
relationship with behavioural disturbances.

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1 **Highlights**

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4

5 (1). Behavioural response was marked by the evolution of 3 parameters

6 (2). Valve closure-duration and valve agitation index increase, valve opening-amplitude decrease

7 (3). Gills were significantly contaminated by PAH but not muscles + ganglia pool

8 (4). Metabolic and cellular processes increased in gills and decreased in muscles + ganglia pool

9 (5). Toxicological impacts seem to be greater in gills than in muscles + ganglia pool

10 **ABSTRACT**

11
12 The use of online remote control for 24/7 behavioural monitoring can play a key role in estimating the
13 environmental status of aquatic ecosystems. Recording the valve activity of bivalve molluscs is a
14 relevant approach in this context. However, a clear understanding of the underlying disturbances
15 associated with behaviour is a key step. In this work, we studied freshwater Asian clams after
16 exposure to crude oil (measured concentration, $167 \pm 28 \mu\text{g}\cdot\text{L}^{-1}$) for three days in a semi-natural
17 environment using outdoor artificial streams. Three complementary approaches to assess and explore
18 disturbances were used: behaviour by high frequency non-invasive (HFNI) valvometry, tissue
19 contamination with polycyclic aromatic hydrocarbons (PAH), and proteomic analysis. Two tissues
20 were targeted: the pool adductor muscles – retractor pedal muscle – cerebral and visceral ganglia,
21 which is the effector of any valve movement and the gills, which are on the frontline during
22 contamination. The behavioural response was marked by an increase in valve closure-duration, a
23 decrease in valve opening-amplitude and an increase in valve agitation index during opening periods.
24 There was no significant PAH accumulation in the muscle plus nervous ganglia pool, contrary to the
25 situation in the gills, although the latter remained in the low range of data available in literature. Major
26 proteomic changes included (i) a slowdown in metabolic and/or cellular processes in muscles plus
27 ganglia pool associated with minor toxicological effect and (ii) an increase of metabolic and/or cellular
28 processes in gills associated with a greater toxicological effect. The nature of the proteomic changes is
29 discussed in terms of unequal PAH distribution and allows to propose a set of explanatory
30 mechanisms to associate behaviour to underlying physiological changes following oil exposure. First,
31 the first tissues facing contaminated water are the inhalant siphon, the mantle edge and the gills. The
32 routine nervous activity in the visceral ganglia should be modified by nervous information originating
33 from these tissues. Second, the nervous activity in the visceral ganglia could be modified by its own
34 specific contamination. Third, a decrease in nervous activity of the cerebral ganglia close to the mouth,
35 including some kind of narcosis, could contribute to a decrease in visceral ganglia activity via a
36 decrease or blockage of the downward neuromodulation by the cerebro-visceral connective. This
37 whole set of events can explain the decrease of metabolic activity in the adductor muscles, contribute
38 to initiate the catch mechanism and then deeply modify the valve behaviour.

39 **KEYWORDS**

40 Biomonitoring

41 Oil spill

42 Proteomics

43 Behavior

44 PAH

45 Asiatic clam

1. INTRODUCTION

Biomonitoring is key in estimating the environmental status of wildlife and their habitats (Goldberg et al., 1978). This is especially true in aquatic ecosystems. Within an environment subject to natural variations and/or deeply disturbed by human activity, recording the behaviour of organisms is a first response tool that requires consolidation to a mature stage. Disrupted behaviour can be used to establish a link between physiological and ecological alterations (Amiard-Triquet, 2009). Today, behavioural ecotoxicology appears reliable in hazard identification, release assessment, exposure assessment, consequence assessment and risk estimation (Hartmann et al., 2016; Pyle and Ford, 2017).

The analysis of valve activity of clams is of particular interest because it allows wireless, 24/7 insight into multiple aspects of the life history of groups of animals over long distances, for 1–3 years (Andrade et al., 2016; Danovaro et al., 2016). Clams are ecologically relevant because most of them are strictly sedentary, sessile, abundant and available year-round. They are filter feeders, which means there is no need to feed them, and using lightweight electrodes and soft cables, one can record artefact-free behaviour readings. Understanding the physiological changes associated with clam-based behavioural monitoring systems is a key step to improving their use for biomonitoring surveys. Hence, proteomic analysis was used in the current study. Protein identification and quantification of individuals subjected to different situations, completed by functional annotations and pathway analysis of proteins, are powerful tools providing access to a global or detailed view of altered biological processes (Campos et al., 2012; Lemos et al., 2009; Van Aggelen et al., 2010). This study was performed on the freshwater bivalve *Corbicula fluminea* in a semi-natural environment within outdoor artificial streams during the first 3 d of crude oil exposure (Miserazzi et al., 2020). Relatively few studies on the effects of crude oil have been carried out in freshwater environments (Bhattacharyya et al., 2003; Green and Trett, 1989; Monteiro et al., 2019). Today, such studies are still in the minority although multiple anthropogenic causes are responsible for oil pollution in freshwaters, including: oil spills, pipeline failure, urban runoff, waste oils or crude oil extraction, etc. (Green and Trett, 1989). Notably, the solubility of crude oil hydrocarbons is generally greater in freshwater than seawater (Eganhouse and Calder, 1975; Green and Trett, 1989; Rossi and Thomas, 1981) and contamination by polycyclic aromatic hydrocarbons (PAH) increases when salinity decreases (Ramachandran et al., 2006; Shukla et al., 2007).

In this study, we first examined key parameters describing the behavioural responses of clams exposed to crude oil by comparison to a reference situation (natural variations only). Thereafter, we focused on the impact of oil in their internal medium by examining tissue that are centrepieces to explain bivalve behaviour, including: (i) the pool adductor muscles-pedal retractor muscles-visceral ganglia-cerebral

82 ganglia and (ii) the gills. We first investigated their respective PAH contamination status and
83 thereafter, explored associated disturbances of major underlying biological-processes by proteomic
84 analysis. Finally, based on this set of indicators, behaviour, PAH contamination and proteomic
85 analysis, we propose a couple of explanatory mechanisms linking behaviour and disturbance of the
86 internal medium in crude oil exposed bivalve molluscs.

87
88 This report is the sister report of a recent work on distinguishable behavioural responses to crude oil in
89 Asian clams *Corbicula fluminea* under semi-natural multiple stress conditions (Miserazzi et al., 2020).
90 In brief, to mimic field conditions and for the purposes of the present experiment, clams were exposed
91 in two outdoor artificial streams that were subject to natural variations and were continuously fed by
92 freshwater from the Gave de Pau river (S.W. France). After a 26-day period, the clams (n = 16 per
93 condition) were separately exposed for an additional period of 10 days to crude oil or natural changes
94 alone. In each stream, a longitudinal survey was performed that involved repeated observations of the
95 same behavioural parameters on the same individuals. The results presented here were obtained by
96 sampling *C. fluminea* exposed for 3 days to crude oil alone or natural changes alone, once the
97 behavioural response curve had flattened. In the presence of crude oil, the behaviour of clams was
98 systematically characterised by an increase in valve-closure duration, a decrease in valve-opening
99 amplitude and an increase in valve agitation index. We present here insights into the nature of the
100 underlying associated proteomic changes.

101 102 103 104 **2. MATERIALS & METHODS**

105 106 **2.1. Clam sampling and maintenance**

107
108 Asian clams *C. fluminea* (height, 24–28 mm) were collected in the northern part of the Parentis -
109 Biscarrosse lake, France. Clams were first maintained in 600 L tanks in an open circuit of freshwater
110 for 10 d at the laboratory (Marine Biological Station, Arcachon, France). During this maintenance
111 period, they were equipped with electromagnets to study behaviour by HFNI valvometry (Andrade et
112 al., 2016). Thereafter, all clams (for behavioural, proteomic and PAH tissue contamination analyses)
113 were introduced in artificial stream facilities of TOTAL (Lacq, southwest France; Bassères and
114 Tramier, 2001).

115 116 **2.2. Artificial stream facility**

118 The artificial stream facility is an outdoor experimental site exposed to natural variations and supplied
119 with freshwater from the Gave de Pau River. The water is derived upstream of a dam, flows by gravity
120 in a flow control system, passes through a living organism “nursery” and is identically distributed in
121 16 experimental streams (length, 40 m; width, 0.5 m; depth, 0.5 m; water depth, 0.25 m) arranged in
122 parallel (Cailleaud et al., 2019; Miserazzi et al., 2020). At the beginning of streams, a piston pump
123 (Prominent) coupled to a high-pressure pump and a shearing valve (Netzsch; Nemo) allows to
124 continuously inject and disperse crude oil in water as micro-droplets. For the requirements of the
125 experiment, in each artificial stream used, quartz and silica sand zones (length, 1 m; depth, 0.1 m)
126 were created to allow the natural burrowing behaviour of clams. The water velocity at clam’s level
127 was $6 \text{ cm}\cdot\text{s}^{-1}$ and the average residence time of water was $\approx 22 \text{ min}$ in a stream.

128 129 **2.3. Experimental protocol**

130
131 The experiment was carried out from October to December 2016. The experiment was divided into
132 three periods with t_0 being the beginning of the exposure period. Period 1 was a 16 d acclimation
133 period (26th October –10th November) plus a 10 d reference period (11th November – 20th November; t_{10}
134 t_{10} – t_{11}). Period 2 (21st November–1st December; t_0 – t_{10}) was the exposure period. The exposure period
135 started on t_0 at 03:15 PM (GMT + 1; 21st November). It ended on t_{10} at 03:15 PM (GMT+1; 1st
136 December). Period 3 (2nd–11th December; t_{11} – t_{20}) was a 10 d post-exposure period. During the
137 experimental period, water temperature varied from 6.6–12.5 °C, O₂ content from 8–13 mg·L⁻¹, pH
138 from 6.5–9.5 and conductivity ≈ 200 –300 $\mu\text{S}\cdot\text{cm}^{-1}$. For the requirements of the experiment, only two
139 artificial streams were used. In one stream, clams were exposed to natural variations only. This is the
140 control (CTRL) condition. In the other stream, clams were exposed to crude oil. This is the crude oil
141 (OIL) condition. We have used non-replicated stream conditions due to logistical problems. For OIL
142 condition, crude oil was continuously injected into the stream ($6.5 \text{ mL}\cdot\text{h}^{-1}$) to obtain a nominal
143 concentration of $400 \mu\text{g}\cdot\text{L}^{-1}$. In each stream, the behaviour of clams was continuously recorded ($n = 16$
144 clams per stream, see section 2.6 for details). These clams were in a first sand zone located at 19–20 m
145 from the crude oil injection. Clams used for tissue sampling were in a second sand zone, located at 29–
146 30 m from the crude oil injection ($n = 5$ clams per stream for proteomic analysis, see section 2.7 for
147 details; $n = 5$ clams per stream for contamination analysis, see section 2.8 for details). The spacing
148 between the two zones made it possible to sample in the second zone without disturbing clams in the
149 first zone (Miserazzi et al., 2020).

150 151 **2.4. Crude oil exposure**

152 153 **2.4.1. Crude oil characteristics**

154 Crude oil was a light oil from the North Sea. Its density was 0.77 (at 15 °C) and dynamic viscosity was
155 1.016 mPa·s (at 15 °C). Saturates, aromatics, resins and asphaltenes (SARA) analysis indicated that
156 the residue fraction (representing 28.2 % of crude oil, compared to 71.8 % for the distillate fraction)
157 contained 84 % saturated hydrocarbons, 15.5 % aromatic hydrocarbons and 0.5 % polar compounds.

158 159 **2.4.2. Analysis of crude oil contamination**

160 Crude oil injection rates were checked at t_1 , t_3 , t_4 , t_7 , t_8 and t_9 . The measured average of crude oil
161 injection rate was $6.6 \pm 0.43 \text{ mL}\cdot\text{h}^{-1}$ (mean \pm 1 SE). Water was sampled at t_1 , t_3 and t_{10} , 35 m
162 downstream of the crude oil injection in the centre of the stream and in the centre of water column in 1
163 L glass bottles (see Fig.1 in Miserazzi et al. (2020) for a detailed diagram with location of the
164 sampling point). Water samples were stabilised with methanol and nitric acid. Total petroleum
165 hydrocarbons (TPH) were extracted with 40 mL of hexane and concentrated under nitrogen flux for
166 analysis by gas chromatography with flame-ionization detection (GC-FID), using an Agilent 7890 B
167 GC system (equipped with a 15 m Agilent CP7491 column) and integration of total peak area between
168 C10 and C40. Quality assurance and quality control (QA/QC) were ensured by the calibration curve
169 method (Shrivastava and Gupta, 2011), the analysis of analytical blanks and the analysis of non-
170 contaminated field samples. The mean measured concentration was $167 \pm 28 \mu\text{g}\cdot\text{L}^{-1}$.

171 172 **2.6. Behaviour analysis**

173 174 **2.6.1. HFNI valvometry**

175 In each stream, the behaviour of 16 clams was continuously assessed by High Frequency Non-
176 Invasive (HFNI) valvometry (Andrade et al., 2016; Tran et al., 2003 for details). Briefly, a pair of
177 lightweight electromagnets (< 1 g) were glued, face-to-face, on each valve. The voltage variation
178 produced by the electromagnetic current generated between the two electromagnets was recorded
179 every 1.6 s for each bivalve. Daily, data were automatically transmitted to a processing unit (Dell 8
180 core processor) based in the Marine Biological Station of Arcachon. Data were automatically handled
181 and processed before daily publication on professional pages of the MolluSCAN *eye* website
182 (<https://molluscan-eye.epoc.u-bordeaux.fr/>).

183 184 **2.6.2. Behavioural parameters**

185 *Hourly valve closure-duration (VCD) and hourly valve opening-amplitude (VOA)*. For each hour and
186 each bivalve, the VCD (below 5 % of valve opening amplitude) and the VOA were reported in
187 percentage. For example, if the bivalve was opened for 1 h, the percentage of VCD was 0 % and if the
188 bivalve was closed for 1 h, the percentage was 100 %. The hourly VOA values 100 % and 0 % were
189 reported from the maximum and minimum values of valve opening, constantly slipping for each

190 bivalve over the previous 6 d. Thereafter, these percentages were determined for the group, averaging
191 16 bivalve values each hour.

192 *Valve agitation index (VAI)*. An index of valve agitation weighted was calculated by dividing the
193 average hourly distance travelled (mm) by electromagnets (i.e. by the valves) by the percentage of
194 hourly VOA (%) for the group of 16 bivalves (Miserazzi et al., 2020).

196 **2.6.3. Behavioural data treatment**

197 Taking into account (i), the beginning of clam exposure to crude oil (21th November; t_0 ; 03:15 PM;
198 GMT + 1), (ii) the time for 50 % of clams to react (465 min or 7 h and 45 min; Miserazzi et al., 2020)
199 and (iii), the sampling time of clams exposed to crude oil for proteomic analyses (24th November; t_3 ;
200 03:00 PM GMT + 1), the bivalve behaviour was analysed from 21th November at 11:00 PM (GMT +
201 1; t_0) to 24th November at 03:00 PM (GMT + 1; t_3). It represents a 64 h exposure period (i.e. 2 days
202 and 16 h) under oil-exposed conditions. Behavioural data were processed with R software (R Core
203 Team, 2016). After verification of normality and homoscedasticity of error terms, comparisons
204 between the two independent variables were investigated using the non-parametric Wilcoxon-Mann-
205 Whitney test. Statistical differences were considered significant at $p < 0.05$.

207 **2.7. Proteomic analysis**

209 **2.7.1. Sampling and dissections of clams**

210 Five clams exposed to crude oil (OIL) and five clams exposed to natural variations only (CTRL) were
211 dissected immediately after collection from artificial streams. Clams in OIL condition were dissected
212 at 03:00 PM (GMT + 1) on t_3 (24th November). Clams in CTRL condition were dissected on the same
213 day at 01:00 PM (GMT + 1). Muscles plus ganglia pool ($n = 5$) and gills ($n = 5$) were sampled and
214 stored at -20 °C for 1 d and then at -80 °C until proteomic analysis was performed.

216 **2.7.2. Sample preparation and protein digestion**

217 Total protein extraction was performed by classical TCA-acetone method. Protein samples were
218 solubilised in Laemmli buffer and 10 µg per sample was deposited onto an SDS-PAGE gel for
219 concentration and cleaning purposes. Separation was stopped once proteins entered the resolving gel.
220 After colloidal blue staining, bands were cut out from the SDS-PAGE gel and subsequently cut in 1
221 mm × 1 mm gel pieces. Gel pieces were destained in 25 mM ammonium bicarbonate 50 % ACN,
222 rinsed twice in ultrapure water and shrunk in ACN for 10 min. After ACN removal, gel pieces were
223 dried at room temperature, covered with trypsin solution (10 ng·µl⁻¹ in 50 mM NH₄HCO₃), rehydrated
224 at 4 °C for 10 min, and finally incubated overnight at 37 °C. Spots were then incubated for 15 min in
225 50 mM NH₄HCO₃ at room temperature with rotary shaking. The supernatant was collected, and an
226 H₂O/ACN/HCOOH (47.5:47.5:5) extraction solution was added onto gel slices for 15 min. The

227 extraction step was repeated twice. Supernatants were pooled and dried in a vacuum centrifuge.
228 Digests were finally solubilised in 0.1 % HCOOH.

230 **2.7.3. nLC-MS/MS analysis**

231 Peptide mixture was analysed on an Ultimate 3000 nanoLC system (Dionex, Amsterdam, The
232 Netherlands) coupled to a Electrospray Orbitrap Fusion™ Lumos™ Tribrid™ Mass Spectrometer
233 (Thermo Fisher Scientific, San Jose, CA). Ten microliters of peptide digests were loaded onto a 300-
234 µm-inner diameter × 5 mm C₁₈ PepMap™ trap column (LC Packings) at a flow rate of 10 µL·min⁻¹.
235 Peptides were eluted from the trap column onto analytical 75-mm id × 50-cm C18 Pep-Map column
236 (LC Packings) with a 4–40 % linear gradient of solvent B in 108 min (solvent A was 0.1 % formic
237 acid and solvent B was 0.1 % formic acid in 80 % ACN). The separation flow rate was set at 300
238 nL·min⁻¹. The mass spectrometer operated in positive ion mode at a 2 kV needle voltage. Data were
239 acquired using Xcalibur 4.1 software in a data-dependent mode. MS scans (*m/z* 375-1500) were
240 recorded at a resolution of $R = 120\,000$ (@ *m/z* 200) and an automatic gain control (AGC) target of 4
241 $\times 10^5$ ions collected within 50 ms. Dynamic exclusion was set to 60 s and top speed fragmentation in
242 higher-energy collision dissociation cell (HCD) mode was performed over a 3 s cycle. MS/MS scans
243 with a target value of 3×10^3 ions were collected in orbitrap (with a resolution of $R = 30\,000$ (@ *m/z*
244 200)) with a maximum fill time of 54 ms. Additionally, only + 2 to + 7 charged ions were selected for
245 fragmentation. Other settings were as follows: no sheath nor auxiliary gas flow, heated capillary
246 temperature, 275 °C; normalised HCD collision energy of 30 % and an isolation width of 1.6 *m/z*.
247 Monoisotopic precursor selection (MIPS) was set to Peptide and an intensity threshold was set to $2.5 \times$
248 10^4 .

250 **2.7.4. Database search and results processing**

251 Data were searched by SEQUEST through Proteome Discoverer 1.4 (Thermo Fisher Scientific Inc.)
252 against a laboratory made protein database (69376 entries). Spectra from peptides higher than 5000 Da
253 or lower than 350 Da were rejected. The search parameters were as follows: mass accuracy of the
254 monoisotopic peptide precursor and peptide fragments set to 10 ppm and 0.6 Da respectively. Only b-
255 and y-ions were considered for mass calculation. Oxidation of methionines (+ 16 Da) and protein N-
256 terminal Acetylation (+ 42 Da) were considered as variable modifications and carbamidomethylation
257 of cysteines (+ 57 Da) as fixed modification. Two missed trypsin cleavages were allowed. Peptide
258 validation was performed using the Percolator algorithm (Käll et al., 2007) and only “high confidence”
259 peptides were retained, corresponding to a 1 % False Positive Rate at the peptide level.

261 **2.7.5. Label-Free Quantitative Data Analysis**

262 Raw LC-MS/MS data from muscles plus ganglia pool and gills were imported in Progenesis QI for
263 Proteomics 2.0 (Nonlinear Dynamics Ltd, Newcastle, U.K) in two separate datasets. Data processing

264 includes the following steps: (i) feature detection, (ii) feature alignment across 10 samples, (iii)
265 volume integration for 2–6 charge-state ions, (iv) normalisation on feature ratio median, (v) import of
266 sequence information, and (vi) calculation of protein abundance (sum of the volume of corresponding
267 peptides). Then, data were processed with Perseus (version 1.6.2.3; Tyanova et al., 2016). Data
268 analysis was partially based on the instructions published by Perseus developers (Tyanova and Cox,
269 2018). Normalised intensities were uploaded in the software, filtered to retain only quantified proteins
270 in at least four out of five samples in at least one group (OIL or CTRL). Thereafter, the remaining
271 missing values were replaced from normal distribution by maintaining the default settings. An
272 additional filtration was performed to remove proteins labelled as “contaminant”. A \log_2
273 transformation was subsequently performed for each sample. Data reliability was assessed by
274 histogram visualisation and multi-scatter plot with Pearson correlation for each sample. For each
275 protein, statistical analysis was performed between OIL (n = 5) and CTRL (n = 5) samples with a
276 student t-test at FDR (False Discovery Rate; q-value) < 0.2 for muscles plus ganglia pool or < 0.25 for
277 gills (Benjamini and Hochberg, 1995). For clustering only, protein abundances were normalised with
278 Z-score. To calculate the protein abundance ratio between both conditions, normalised \log_2 intensities
279 (before Z-scoring) was averaged for OIL (n = 5) and CTRL samples (n = 5). Fold change (FC) was
280 obtained by the $2^{\lfloor \log_2(\text{OIL mean}) - \log_2(\text{CTRL mean}) \rfloor}$ operation. To determine which pathways were involved by
281 significant proteins, data were processed with Cytoscape software (version 3.7.1; Shannon et al., 2003)
282 and ClueGO (version 2.5.4; Bindea et al., 2009) plus CluePedia (version 1.5.4; Bindea et al., 2013)
283 plug-ins. Analyses were carried out independently for both tissues and protein clusters (i.e. proteins
284 with highest or lowest abundance). Gene identifiers in the oyster *Crassostrea gigas*, corresponding to
285 selected proteins, were uploaded in the software instead of the database on *Corbicula fluminea* (Chen
286 et al., 2013). We used the *C. gigas* database because it is considered as the most complete reference
287 transcriptome today in aquatic invertebrates. It was realized using embryo-larval development stages
288 and adult organs as well and a large number of environmental stressors (Rivière et al., 2015; Zhang et
289 al., 2012). Gene Ontology (GO) databases for Biological Processes (BP) were applied for gene
290 annotation. All enriched terms were selected using the right-sided hypergeometric test with
291 Benjamini-Hochberg (BH) adjustment at $p < 0.05$. Term fusion was activated for all protein clusters.
292 The network was designed with a kappa score of at least 0.4. The ClueGO Layout was applied and
293 manually adjusted in a minimal way.

294 295 **2.8. Clams contamination**

296 297 **2.8.1. Sampling and dissections of clams**

298 Sampling and dissections of clams for analysis of tissue contamination were performed on t_{10} (1st
299 December; \approx 02:00 PM; GMT + 1). Five clams exposed to crude oil (OIL) and five clams exposed to

300 natural variations only (CTRL) were dissected immediately after collection from artificial streams.
301 Gills (n = 5) and muscles plus ganglia pool (n = 5) were stored at -20 °C until analysis was performed.
302 Tissue contamination levels were analysed at t₁₀ for analytical reasons (low tissue mass and low
303 expected concentrations levels). However, based on literature data, it was assumed that contamination
304 levels by PAH at t₁₀ overestimate the contamination level at t₃ (Foster et al., 1987; Liu et al., 2014;
305 Mason, 1988; Pruell et al., 1986; Stegeman and Teal, 1973; van Haren et al., 1994).

307 **2.8.2. PAH analysis**

308 In brief, 21 PAH (acenaphthylene, acenaphthene, anthracene, benzo(a)anthracene, benzothiophene,
309 benzo(b)fluoranthene, benzo(g,h,i)perylene, benzo(k)fluoranthene, benzo(e)pyrene, benzo(a)pyrene,
310 biphenyl, chrysene, dibenzo(a,h)anthracene, dibenzothiophene, fluoranthene, fluorene, indeno(1,2,3-
311 cd)pyrene, naphthalene, perylene, phenanthrene, and pyrene), among which, 16 listed as priority
312 pollutants by the US EPA (United States Environmental Protection Agency; underlined in the list)
313 were analysed in muscles plus ganglia pool (n = 5) and gills (n = 5) of five clams per condition
314 (Miserazzi et al., 2020). The analyses were performed by stir bar sorptive extraction-thermal
315 desorption-gas chromatography-tandem mass spectrometry (SBSE-GC-MS/MS) as described in
316 Lacroix et al. (2014). Analytes were quantified relative to deuterated compounds using a calibration
317 curve ranging from 0.01 ng to 10 ng per bar. Two compounds, benzo(b)fluoranthene and
318 benzo(k)fluoranthene, were quantified as a sum named benzo(b+k)fluoranthene due to poor resolution.
319 The limits of quantification (LOQ) were calculated by the calibration curve method (Shrivastava and
320 Gupta, 2011), and the limit of detection (LOD) was estimated by dividing the LOQ by 3.

322 **2.8.3. Statistics**

323 Tissue contamination data were processed with R software (R Core Team, 2016). After verification of
324 normality and homoscedasticity of error terms, comparisons between the two independent variables
325 were investigated using the non-parametric Wilcoxon-Mann-Whitney test. Statistical differences were
326 considered significant at p < 0.05.

329 **3. RESULTS**

330
331
332 No mortality occurred during the entire experiment (acclimation period, reference period, exposure
333 period and post-exposure period).

335 **3.1. Behavioural disturbances and tissue contamination**

336
337 The behaviour of clams was deeply modified by the presence of crude oil. It was characterised by
338 three major changes: an increase of valve closure duration, VCD (Fig. 1A), a decrease of valve
339 opening amplitude, VOA (Fig. 1B) and an increase of valve agitation index, VAI (Fig. 1C). The
340 hourly VCD was 9.5 times greater in the presence of crude oil (Fig. 1A). The hourly VOA was 5.3
341 times lesser (Fig. 1B) and the hourly VAI was 2.8 times greater (Fig. 1C). The level of contamination
342 in muscles plus ganglia pool and in gills after a 10 d exposure period is presented in Fig. 1D. All
343 values are in the low range. The median contamination level in muscles plus ganglia pool was not
344 different between exposed and control clams (6.1 and 7.4 ng·g⁻¹ ww). In contrast, it was different in
345 gills. It was 70.2 ng·g⁻¹ ww in gills of OIL clams and 18.9 ng·g⁻¹ in CTRL clams.

346 347 **3.2. Analysis of proteomic changes**

348
349 A total of 1860 proteins were quantified in muscles plus ganglia pool and 5469 in gills. Among them,
350 in the muscles plus ganglia pool, 139 showed a significant difference between the both conditions
351 (7.5 %; FDR < 0.2) and 144 in gills (2.6 %; FDR < 0.25). As an exploratory analysis, Principal
352 Component Analyses (PCA) was applied to significantly changing proteins. PCA showed a clear
353 individualisation according to the experimental conditions, which explained 81.3 % (Component 1,
354 74.2 % + Component 2, 7.1 %) variability in the muscles plus ganglia pool (Fig. 2A) and 80 %
355 (Component 1, 74 % + Component 2, 6 %) in the gills (Fig. 2B). Thereafter, we turned to functional
356 analyses (hierarchical clustering and corresponding pathways analysis) to explore observed differences
357 (Figs. 3A, 3B, 3C and 4A, 4B, 4C). Firstly, in agreement with the PCA, upper vertical clustering
358 shows a clear segregation between crude oil and control conditions (OIL, five left-side columns;
359 CTRL, five right-side columns) in muscles plus ganglia pool (Fig. 3A) and in gills (Fig. 4A). In
360 hierarchical clustering, each horizontal line represents one differentially expressed protein. In both
361 tissues, horizontal hierarchical clustering revealed two major row clusters: in the upper ones, proteins
362 were less abundant in the presence of crude oil and in the lower one, proteins were more abundant in
363 the presence of crude oil. Fig. 3A shows that in muscles plus ganglia pool, 90 proteins (65 % of the
364 total) were less abundant and 49 proteins (35 % of the total) were more abundant in crude-oil exposed
365 clams. In contrast, in gills (Fig. 4A), only 59 proteins (i.e. 41 % of the total) were less abundant in
366 crude-oil exposed clams while 85 proteins (59 % of the total) were more abundant. Thereafter, a
367 functional analysis was performed for each row cluster within the two tissues (Figs. 3B, 3C and 4B,
368 4C). In muscles plus ganglia pool, biological processes associated to less abundant proteins (Fig. 3B)
369 were relatively numerous. There were essentially related to metabolic and/or cellular activities namely
370 generation of precursor metabolites and energy, small molecule metabolic process, organonitrogen
371 compound metabolic process, cellular ketone metabolic process, alpha-ketoacid dehydrogenase
372 activity, proteolysis and glutamate biosynthetic process. GO enrichment analysis also indicated that

373 this set of proteins was associated with the biological regulation of neurotransmitter levels. The most
374 abundant proteins were statistically represented through the categorical term system process, which
375 makes it possible to establish a bridge between regulation of muscle contraction and nervous system
376 through the term synapse assembly (Fig. 3C). The less abundant proteins in oil-contaminated gills
377 were statistically enriched only through two biological processes, negative regulation of endopeptidase
378 activity and regulation of hydrolase activity (Fig. 4B). The functional network associated to proteins
379 more abundant in contaminated gills was relatively richer and the GO terms were also mainly related
380 to metabolic and/or cellular processes (Fig. 4C). More precisely, GO terms were related to cell
381 signalling pathway with activation of MAPKK activity, to energy or nucleotide metabolism with
382 tricarboxylic acid cycle, GTP metabolic process or nucleoside diphosphate phosphorylation. Finally,
383 GO terms were also related to cell signal transduction with adenylate cyclase-modulating G-protein
384 coupled receptor signalling pathway and to protein processing with the term protein polymerization.

387 **4. DISCUSSION**

388
389
390 In the current study, we used three complementary approaches to further increase analytical and
391 empirical understanding of crude oil-induced behaviour in the Asian clam *C. fluminea*: behaviour,
392 PAH contamination and proteomics. The behavioural response of clams exposed to crude oil was
393 unequivocal. It was characterised by a significant increase of VCD, a significant decrease of VOA and
394 a significant increase of VAI (Figs. 1A, 1B, 1C). We sought, through a proteomic analysis focused on
395 two target tissues (the adductor muscles – pedal retractor muscles – visceral ganglia – cerebral ganglia
396 pool and the gills), to explore the underlying disturbances that occur in the internal medium of
397 exposed clams. To complete the analysis, we used data on PAH accumulation that showed significant
398 contamination in gills but no contamination in muscles plus ganglia pool. Importantly, these
399 contamination levels were in the low range of literature data (Webster et al., 2006). The major
400 proteomic changes included a decrease of metabolic and/or cellular processes and nervous functions in
401 muscles plus ganglia pool with a minor toxicological impact, and an increase of metabolic and/or
402 cellular processes in gills with a higher toxicological impact.

403 404 **4.1 Non-replicated stream conditions**

405
406 In this work, we have used non-replicated stream conditions due to logistical problems but we
407 individually analysed five Asian clams exposed to crude oil in one stream and five other individuals
408 exposed to control conditions in another stream. A PCA analysis showed first a clear individualisation

409 between streams (Figs. 2A and 2B). Then we turned to functional analysis and upper vertical
410 clustering that also showed a clear segregation between crude oil and control streams (Figs. 3A and
411 4A). Thus, despite the use of only one stream per treatment, that is a standard situation in field
412 sampling, our data set shows the amplitude of individual variations and strongly supports significant
413 differences between oil and control conditions.

414 415 **4.2. The target tissues**

416
417 We focused our interest on the muscles plus ganglia pool and the gills for evaluating their
418 functional role in setting out Asian clam's behaviour.

419 The pool was composed of the anterior and posterior adductor muscles (which is a mix of striated
420 and smooth muscles), anterior and posterior pedal retractor muscles, paired cerebral ganglia, paired
421 visceral ganglia and its innervations to muscles (Kraemer and Lott, 1977; Britton and Morton 1982).
422 In *C. fluminea*, cerebral ganglia and visceral ganglia are two of the three major ganglia in the nervous
423 system. Cerebral ganglia lie flat against the anterior pedal muscles while visceral ganglia nestle in
424 posterior pedal muscles. Visceral ganglia innervate gills through a gill nerve while cerebral ganglia are
425 connected to the digestive system (oesophagus, stomach, intestine, etc.) and pedal ganglia. The
426 visceral and cerebral ganglia are also interconnected with each other. In bivalves, adductor muscles are
427 mechanical effectors of valve movements with the elastic ligament situated along the edge of the hinge
428 between the valves. When these muscles are impacted, the valve activity is expected to be modified
429 and *vice versa*. The fast muscle (or striated muscle) is responsible for rapid contractions, while the
430 slow muscle (or smooth muscle) is responsible for the state of catch, which is the maintenance of
431 valve closure for long periods with minimal energy demand (Sun et al., 2018).

432 Gills were sampled because they are, together with the inhalant siphon and the mantle edge, on the
433 frontline during any contamination. Gill tissue is particularly complex, and composed of respiratory
434 exchange tissue, ion exchange tissue, immune tissue, ciliary and muscles for water movements and
435 feeding purposes (Britton and Morton, 1982; Gainey et al., 2003; Medler and Silverman, 2001;
436 Mommsen, 1984). It is richly innervated via a branchial nerve originating from the visceral ganglia
437 (Britton and Morton, 1982; Kraemer and Lott, 1977).

438 439 **4.3. Disturbances in muscles + ganglia pool**

440 441 **4.3.1. Proteins with a lower abundance**

442 *Metabolic activity.* Within the muscles plus ganglia pool (Fig. 3), proteins with a lower abundance
443 were a majority. These proteins were mainly involved in metabolic processes. This strongly suggests a
444 decrease in metabolic activity, which is consistent with the change in behavioural activity. Indeed, in
445 exposed clams, there was an increase in valve closing duration and thus an increase in the catch

446 mechanism. In bivalves, catch is a passive state of the smooth muscle that allow valves to remain
447 closed for long periods with minimal energy consumption (Galler, 2008; Yamada et al., 2013). The
448 valve opening amplitude was also lower, which is consistent with the decline in valve activity.
449 Altogether, this behaviour represents probably a defence mechanism to limit exposure of the inner soft
450 tissues. With regard to the VAI, exposed clams were more agitated when opened. However, valve
451 closing duration was significantly greater in exposed clams, suggesting that transient valve agitation
452 behaviour has a limited weight in the overall energetic balance of valve activity.

453 *Neurotransmitters.* Taking a close look at the lower abundant proteins impacted in the muscles plus
454 ganglia pool, it can be seen that some elements were linked to the metabolism of glutamate. Glutamate
455 is a major metabolic hub involved in diverse processes (Walker and van der Donk, 2016) and, is an
456 excitatory neurotransmitter, whose deregulation can lead to some toxicity in the nervous system (Choi,
457 1988; Wu and Wang, 2010). The network analysis of lower abundant proteins highlights the disruption
458 of the "glutamate biosynthetic process" and, interestingly, glutamate synthase (FC, 0.018; FDR, 0.082)
459 was found among the top five of less abundant proteins in this tissue pool. Its decrease decreases
460 glutamate synthesis (from 2-oxoglutarate, an intermediate of the tricarboxylic acid cycle), which is in
461 agreement with results from Müller et al. (2018) on oysters. These authors observed a decrease in
462 glutamate 4 receptor abundance in gill microsomes of oysters *Crassostrea brasiliana* exposed during
463 24 h to water contaminated by diesel fuel.

464 With regard to the regulation of neurotransmitter levels, we identified two other proteins through
465 network analysis of lower abundant proteins: synaptotagmin-1 (FC, 0.12; FDR, 0.187) and 4-
466 aminobutyrate aminotransferase or gamma-aminobutyric acid (GABA) transaminase (FC, 0.08; FDR,
467 0.161). Synaptotagmin-1 plays an important role in neurotransmission by being a major sensor of Ca^{2+}
468 within excitatory synapses (Brose et al., 1992; Chapman, 2002). A decrease of synaptotagmin-1
469 suggests a decrease of excitatory synapse activity. This protein is also implicated, but to a lesser
470 extent, in the transmission of the nervous message within GABAergic neurons (Kerr et al., 2008).
471 GABA transaminase allows the synthesis of glutamate from 2-oxoglutarate - an intermediate of the
472 tricarboxylic acid cycle - and from GABA. Its lower abundance is consistent with the lower abundance
473 of glutamate synthase, thus supporting the hypothesis of a decrease in glutamate synthesis. Finally,
474 this protein is a key enzyme in GABA catabolism (Walls et al., 2015), a neurotransmitter that inhibits
475 the nervous system of vertebrates and invertebrates (Lunt, 1991; Miller, 2019). We suggest that
476 inhibition of this enzyme, which leads to increased levels of GABA (inhibitor neurotransmitter; Abel
477 and Kohli, 1999), and inhibition of glutamate synthase, responsible for glutamate synthesis (excitatory
478 neurotransmitter), should contribute to the decrease in clam's activity in the presence of crude oil.
479 Indeed, whether in vertebrates or invertebrates, GABA is able to regulate locomotor patterns by
480 modulating underlying neural circuits (Shen et al., 2016). In addition, the change in abundance of the
481 enzyme GABA transaminase in the brain could cause mitochondrial muscle disruption, independent of
482 the GABAergic system. Besse et al. (2015) showed that a deficiency of this enzyme in the brain can

483 cause mitochondrial DNA (mtDNA) depletion, a syndrome associated with various symptoms
484 including muscle weakness and hypotonia. Note, all this forms a coherent whole with the tissue pool
485 analysed and the decrease in behavioural activity.

486 *Mitochondrial process.* The most impacted protein, in the muscles plus ganglia pool, is the ClpX
487 protein (FC, 0.001; FDR, 0.024). This protein belongs to the AAA+ superfamily (ATPases Associated
488 with various cellular Activities), a family of proteins that is found in all kingdoms of life and
489 participates in a wide range of biological processes (Ammelburg et al., 2006; Ogura and Wilkinson,
490 2001; Sysoeva, 2016). More specifically, ClpX protein is involved in dependent ATP proteolysis by
491 being a component of Clp proteases (Baker and Sauer, 2012). It is involved in mitochondrial processes
492 by affecting the distribution and morphology of mtDNA nucleoids (Bogenhagen et al., 2008;
493 Kasashima et al., 2012; Ogura and Wilkinson, 2001). Its lower abundance highlights an effect on
494 proteolysis, in a coherent way with the network analysis of the lower abundant proteins in the muscles
495 plus ganglia pool as well as a potential mitochondrial dysfunction due to the stress generated on the
496 mtADN. In turn, this mitochondrial dysfunction can be apprehended by network analysis, which
497 shows a decrease in the "generation of precursor metabolites and energy". The three proteins
498 associated with this process within the network were electron transfer flavoprotein-ubiquinone
499 oxidoreductase (FC, 0.689; FDR, 0.147), an enzyme from the respiratory electron transport chain,
500 succinyl-CoA ligase[GDP-forming] subunit beta (FC, 0.06; FDR, 0.104), a tricarboxylic acid cycle
501 enzyme, and xanthine dehydrogenase (FC, 0.107; FDR, 0.168), an enzyme leading to the generation of
502 reactive oxygen species (ROS). A lower abundance of xanthine dehydrogenase has already been
503 shown in gills of oysters *Crassostrea gigas* exposed to various stresses: thermal stress, hypo-salinity
504 and aerial exposure (Zhang et al., 2015). The lower expression of these three proteins suggests a
505 decrease in mitochondrial activity, a decrease in number of mitochondria, or both simultaneously.
506 Recently, several studies highlighted the effect of crude oil on mitochondria, particularly concerning
507 their role in aerobic metabolism (Johansen and Esbaugh, 2019; Kirby et al., 2019; Salazar-Coria et al.,
508 2019). Following a 24 h exposure period to crude oil, Kirby et al. (2019) proposed that crude oil may
509 alter the mitochondrial respiratory chain, thus the ATP supply of myocardial cells in mahi-mahi
510 (*Coryphaena hippurus*). This dysfunction was also implied in the impairment of cardiac muscle
511 function observed in the presence of oil in fish (Cox et al., 2017; Johansen and Esbaugh, 2019; Kirby
512 et al., 2019; Nelson et al., 2016; Nelson et al., 2017;). We suggest that the dysfunction of
513 mitochondrial processes within the muscles plus ganglia pool could be at the origin of the decrease, or
514 at least, the alteration in muscular and nervous function of clams, and therefore, by extension, in their
515 behavioural activity.

516 517 **4.3.2. Proteins with a higher abundance**

518 *Neuromuscular synapses.* The detail of higher abundant proteins impacted in the muscles plus ganglia
519 pool shows some disturbances in the nervous system through the "synapse assembly" process as well

520 as a disturbance of the muscular system with the "regulation of muscle contraction" processes. The
521 alteration of the "synapse assembly" process indicates a disruption in the formation of synapses, thus
522 in the potential number of mature and functional synapses. Disruption of muscle contraction involved
523 the disruption of any process able to modulate frequency and/or the extent of muscle activity. The
524 protein linking the synapse assembly process and the system process was neurexin 1 α . Neurexins are
525 transmembrane proteins located mainly at synaptic extremities of neurons and are involved in
526 neurotransmission and formation of synapses (Reissner et al., 2013). Kang et al. (2007) found that α -
527 neurexins promote recruitment and stabilisation of GABAergic synapses. This again supports the role
528 of GABA activity in the global increase of valve closing duration in oil-contaminated clams.
529 Concerning the regulation of muscle contraction, the protein involved was troponin T of skeletal
530 muscle. In the adductor muscle of bivalve molluscs, striated muscle is responsible for rapid valve
531 contractions (Chantler, 2006; Millman, 1967; Sun et al., 2018). Within this tissue, troponin T will bind
532 troponin I and troponin C to tropomyosin (Farah and Reinach, 1995; Tanaka et al., 2008; Zot and
533 Potter, 1987). To conclude, the whole data set above showed that both nervous and muscle tissues
534 were impacted by the inspired crude oil.

535 536 **4.3.3. Stress markers and oxidative stress**

537 In their meta-analysis of the literature, Anderson et al. (2015) identified the top five genes most
538 commonly affected by environmental stressors. In the category "intracellular stress and
539 detoxification", we find genes encoding superoxide dismutase, metallothioneins, glutathione-S-
540 transferases, heat shock protein (HSP) 70's and glutathione peroxidases. Among these key stress
541 markers, in the muscles plus ganglia pool, three were present: a chaperone protein, the heat shock
542 cognate 71 kDa protein (FC, 1.58; FDR, 0.053) and two enzymes involved in defence against
543 oxidative stress, superoxide dismutase Cu/Zn (FC, 16.5; FDR, 0.151) and glutathione peroxidase (FC,
544 0.119; FDR, 0.196). Cu/Zn superoxide dismutase catalyses the transformation of the superoxide anion
545 into hydrogen peroxide, outside of the mitochondrial matrix. Its higher abundance would indicate an
546 excess of the superoxide anion (despite global decrease in metabolic activity). Glutathione peroxidase
547 catalyses the transformation of hydrogen peroxide into water within the mitochondrial matrix. Its
548 lower abundance consequently indicates either a deficit in hydrogen peroxide within the mitochondrial
549 matrix or a failure in reducing hydrogen peroxide toxicity. Regardless, the respective abundance of
550 these two enzymes suggests an excess of hydrogen peroxide, a key metabolite of oxidative stress.
551 Hydrogen peroxide is a messenger of the redox signal at the enzymatic and transcriptional levels (Sies,
552 2017). It is able to pass through biological membranes, affect other molecules (lipids, proteins, DNA),
553 whether near or distant from its original site, and interfere with other biochemical reactions (Tomanek,
554 2015). In a manner coherent with the literature (Baussant et al., 2009; Boutet et al., 2004; Jiang et al.,
555 2017; López-Landavery et al., 2019 for examples in bivalves), there was a disruption of enzymes
556 involved in defence against oxidative stress in response to crude oil exposure.

4.3.4. A specific presence of PAH in the nervous system?

Multiple enzymatic responses to oxidative stress were shown in the muscle plus ganglia pool despite an absence of local PAH accumulation in this specific tissue set. However, the non-significant accumulation in the muscles plus ganglia pool (Fig. 1D) should not be directly interpreted as an absence in ganglia. Indeed, the fraction of nervous tissue was minor in our samples. An unequal PAH distribution between muscle and ganglia could occur. Lipid content is not the only biological determinant governing PAH accumulation in tissue (Frapiccini et al., 2018), but the nervous tissue is made of $\approx 80\%$ of lipids in invertebrates and vertebrates (McColl and Rossiter, 1951), while there are only $\approx 4\%$ of lipids in the adductor muscles in bivalve molluscs (Dongre and Sonwane, 2014). We suggest in bivalves, that the nervous tissue could be a PAH sink as in fishes. In the latter, the brain accumulates largest quantities of PAH by comparison to muscles (Wu et al., 2012; Xu et al., 2011). However, considering the low amount of tissue available for PAH analysis, the latter remains an open challenge in small bivalve molluscs such as *Corbicula*.

4.4. Disturbances in gills

In gills, where accumulation of PAH was highly significant, there was a deregulation of two HSPs 70's only: 12A (FC, 1.23; FDR, 0.114) and 12B (FC, 0.079; FDR, 0.153), that are among the top five most commonly affected genes in the category "intracellular stress and detoxification" listed by Anderson et al. (2015). HSP plays a key role in protein regulation and were considered proteins produced or deregulated by all cellular organisms in a stressful situation (Roberts et al., 2010). However, network analysis of higher abundant proteins also shows the activation of MAPKK activity (mitogen-activated protein kinase kinase; Fig. 4). MAPKK is part of the MAPK (mitogen-activated protein kinase) phosphorelay system, a group of serine/threonine kinases that participate in gene expression regulation, mitosis, movement, metabolism and apoptosis (Johnson and Lapadat, 2002). Multiple studies have highlighted the disruption of this pathway in the presence of hydrocarbons or other pollutants in aquatic organisms (Burlando et al., 2006; Châtel et al., 2011; Won et al., 2016). In addition, we were able to notice the presence of other proteins reflecting (i) the disruption of the cell cycle, (ii) the disruption of the apoptotic mechanism or (iii) the accumulation of potential DNA damage. RecQ1, a protein that plays an important role in cellular DNA metabolism (Wu and Brosh, 2010) by being involved in maintaining genome stability, was lower abundant in exposed clams (FC, 0.064; FDR, 0.248). There was also the lower abundance (FC, 0.041; FDR, 0.193) of Cdc5-like protein, a DNA-binding protein involved in the regulation of the mitotic cell cycle and in DNA damage response (Bernstein and Coughlin, 1998; Mu et al., 2014). Lastly, caspase 7, a protease involved in programmed cell death and the inflammatory process (Lamkanfi and Kanneganti, 2010),

594 was lower abundant (FC, 0.427; FDR, 0.223). Finally, there was the higher abundance of the tumour
595 protein D54 (FC, 1.90; FDR, 0.218) which belongs to the protein family D52 (Nourse et al., 1998).
596 Proteins in this family, in addition to being involved or suspected of being involved in many cellular
597 processes, are frequently overexpressed in multiple cancers where they are thought to play a role in
598 cell proliferation and thus promote tumour progression (Boutros et al., 2004). In addition, network
599 analysis of higher abundant proteins in gills show a general increase in energy and nucleotide
600 metabolism, an increase in transduction of cellular signals and an increase in protein processing. The
601 analysis thus shows an increase in cellular activity in gills including a substantial mobilisation of
602 energy resources. This increase may be necessary to promote cell maintenance in response to crude oil
603 exposure and PAH accumulation in gills of clams after a 3 d exposure period. This general metabolic
604 change could also promote or support the potential carcinogenesis process initiated in this tissue (Lunt
605 and Fendt, 2018).

607 **4.5. Mechanisms that may explain the behavioural response to crude oil**

608

609 To conclude and propose a link between behavioural observations, proteome modifications and
610 PAH contamination reported in this study, we suggest two mechanisms or scenarios to relate
611 behavioural disturbances to those of the internal medium (Figs. 5A, 5B, 5C). Before that, one must
612 recall that in clams, the first tissue facing contaminated water are the inhalant siphon and to a lesser
613 extent the rest of the mantle edge (Fig. 5A). These areas have many sensory papillae or tentacles
614 capable of tasting the ambient water and thus leading to the valve closure (Britton and Morton, 1982).
615 In the first scenario (Fig. 5B), the routine activity of visceral ganglia would be disrupted first by
616 nervous information generated by the contact of the contaminants with sensory papillae of the inhaling
617 siphon and/or mantle edges and then the contaminated gills. In a second step, the nervous activity of
618 visceral ganglia could be disrupted by its own specific contamination. According to proteomic
619 analyses within the muscles plus ganglia pool, the list of disturbances includes a decrease of nervous
620 and metabolic activity. We propose that the decrease of nervous activity in visceral ganglia
621 participated to decrease metabolic activity in muscles, to decrease valve activity and contributed to the
622 catch mechanism. In the second scenario (Fig. 5C), we include a neuromodulatory role for cerebral
623 ganglia. The nervous activity of the cerebral ganglia could also be disrupted by its own specific
624 contamination. Moreover, cerebral ganglia have a special location near the mouth, which exposes
625 them more to hydrocarbon compounds. Indeed, all contaminated mucus collected on gills converges to
626 the mouth and then to the cerebral ganglia area, which could be affected by a passive diffusion of
627 hydrocarbons. The decrease in nervous activity of the cerebral ganglia, including some kind of
628 narcosis, would cause an additional decrease in nervous activity of visceral ganglia via a decrease or
629 blockage of neuromodulation (mechanism reviewed by Hamood and Marder, 2014). This would

630 strengthen the effects already presented in the first scenario on visceral ganglia, muscles and thus
631 valve activity.

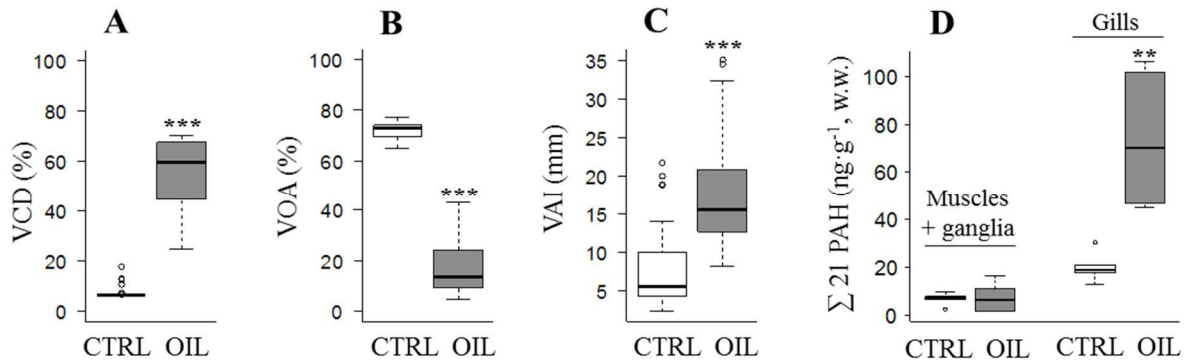
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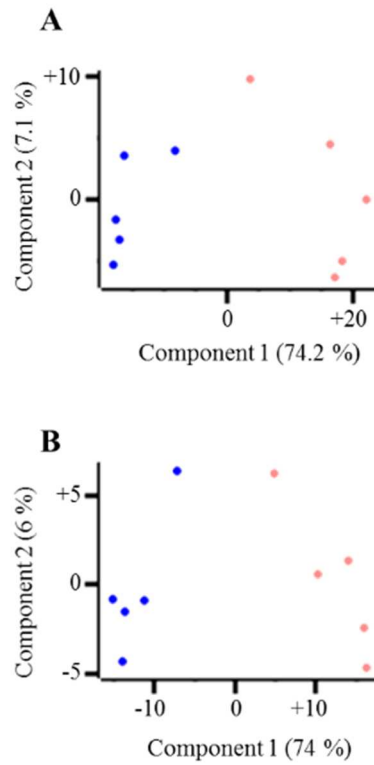
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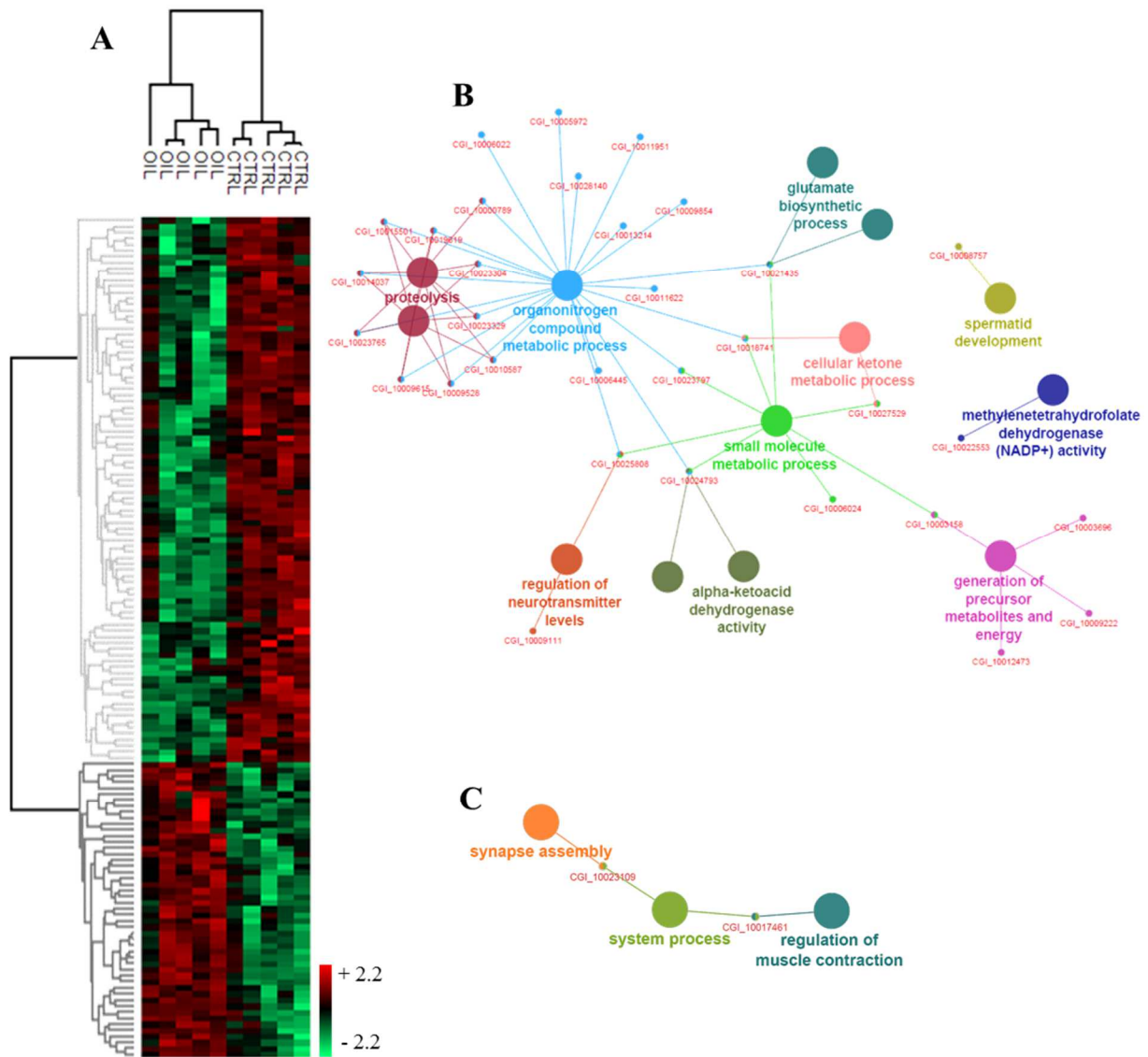
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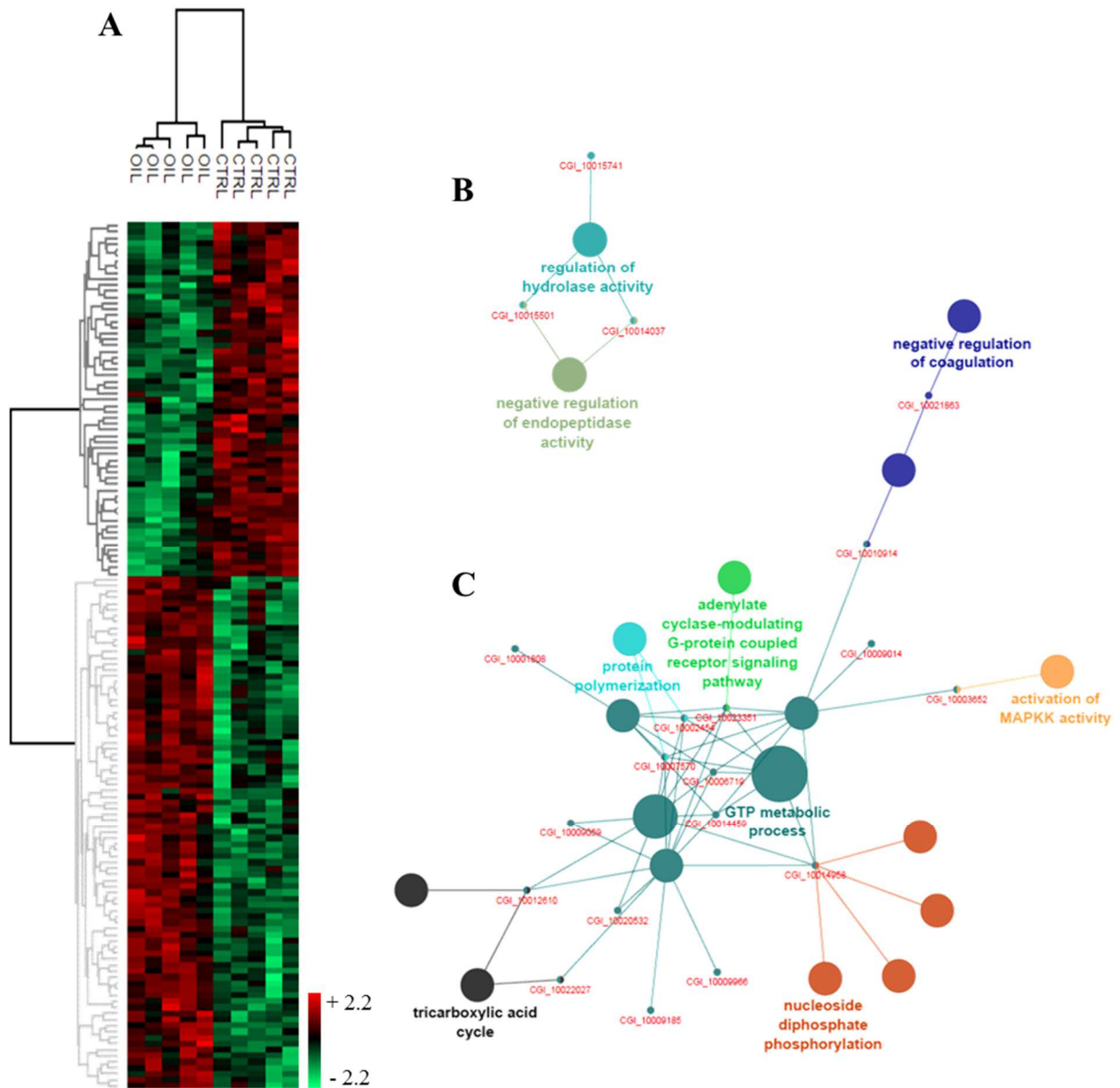
Fig. 1. Behavioural analysis and tissue contamination. (A) Mean hourly valve closure-duration VCD (%; n = 64 for 64 h), (B) mean hourly valve opening-amplitude VOA (%; n = 64 for 64 h), (C) mean hourly valve agitation index VAI (mm; n = 64 for 64 h) of 16 clams by condition, (D) sum of individual concentrations of 21 PAH analysed in muscles plus ganglia pool and gills after 10 d of exposure (ng·g⁻¹, wet weight; n = 5 clams per condition. Redrawn from Fig. 7 in Miserazzi et al., 2020). Data are given as a boxplot. The top line, middle line and lower line of the box indicate upper quartile, median and lower quartile respectively. The ends of the whiskers indicate the lowest and highest values. The dots are outliers. **, p < 0.01; ***, p < 0.001; Wilcoxon-Mann-Whitney. CTRL, clams under control condition; OIL, clams exposed to crude oil.



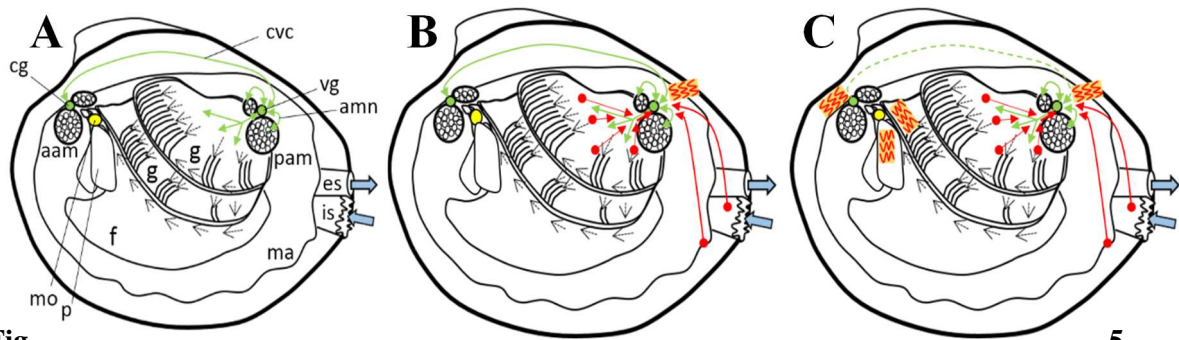
666 **Fig. 2. Principal Component Analysis.** PCA performed with significant proteins in (A), muscles plus
 667 ganglia pool and (B) gills. It clearly separates the five OIL clams (exposed to crude oil; pink circles,
 668 on the right) from the five CTRL clams (control conditions; blue circles, on the left).



669 **Fig. 3. Differential proteomic analysis performed in muscles plus ganglia pool of crude oil**
 670 **exposed clams vs control.** (A) Hierarchical clustering with significant proteins (FDR < 0.2).
 671 Normalized z-scored protein abundance are depicted in red (relatively higher abundance) and green
 672 (relatively lower abundance). Two row clusters of protein profiles are highlighted in light and dark
 673 grey. (B) Functional network of the proteins with relatively lower abundance in OIL clams ($p < 0.05$).
 674 (C) Functional network of proteins with relatively higher abundance in OIL clams ($p < 0.05$). The
 675 dashed line separates the analyses between lower and higher abundances.



676 **Fig. 4. Differential proteomic analysis in gills of crude oil exposed clams vs control.** (A)
 677 Hierarchical clustering with significant proteins (FDR < 0.25). Normalized z-scored protein
 678 abundance are depicted in red (relatively higher abundance) and green (relatively lower abundance).
 679 Two row clusters of protein profiles are highlighted in light and dark grey. (B) Functional network of
 680 proteins with a relatively lower abundance in OIL clams ($p < 0.05$). (C) Functional network of proteins
 681 with a relatively higher abundance in OIL clams ($p < 0.05$). The dashed line separates analyses
 682 between lower and higher abundances.
 683



684 **Fig.** 5.

685 **Schematic diagrams that potentially explain behavioural changes in crude oil exposed Asian**
 686 **clams.** (A) Background information on the anatomy of *C. fluminea* and two nonexclusive disturbance
 687 scenarios (B and C) explaining behavioural changes reported in this work, see text. aam, anterior
 688 adductor muscle; cg, cerebral ganglia; cvc, cerebro-visceral commissure; es, exhalant siphon; f, foot;
 689 g, gills; is, inhalant siphon; ma, mantle; mo, mouth; p, labial palp; pam, posterior adductor muscle; vg,
 690 visceral ganglia. Green arrows, efferent nerves from visceral ganglion to adductor muscles, gills and
 691 between visceral and cerebral ganglia. Red arrows, afferent nerves from inhalant siphon, mantle edge
 692 and gills to visceral ganglia. Wavy red lines represent the presence of hydrocarbon compounds
 693 directly acting on ganglia.

694 Abel, M.S., Kohli, N., 1999. GABA-transaminase antisense oligodeoxynucleotide modulates cocaine- and
695 pentylenetetrazol-induced seizures in mice. *Metab Brain Dis* 14, 253–263.

696 Amiard-Triquet, C., 2009. Behavioral Disturbances: The Missing Link between Sub-Organismal and Supra-
697 Organismal Responses to Stress? Prospects Based on Aquatic Research. *Human and Ecological Risk*
698 Assessment: An International Journal 15, 87–110. <https://doi.org/10.1080/10807030802615543>

699 Ammelburg, M., Frickey, T., Lupas, A.N., 2006. Classification of AAA+ proteins. *J. Struct. Biol.* 156, 2–
700 11. <https://doi.org/10.1016/j.jsb.2006.05.002>

701 Anderson, K., Taylor, D.A., Thompson, E.L., Melwani, A.R., Nair, S.V., Raftos, D.A., 2015. Meta-
702 Analysis of Studies Using Suppression Subtractive Hybridization and Microarrays to Investigate the
703 Effects of Environmental Stress on Gene Transcription in Oysters. *PLoS One* 10.
704 <https://doi.org/10.1371/journal.pone.0118839>

705 Andrade, H., Massabuau, J.-C., Cochrane, S., Ciret, P., Tran, D., Sow, M., Camus, L., 2016. High
706 frequency non-invasive (HFNI) bio-sensors as a potential tool for marine monitoring and assessments.
707 *Frontiers in Marine Science*. <https://doi.org/10.3389/fmars.2016.00187>

708 Baker, T.A., Sauer, R.T., 2012. ClpXP, an ATP-powered unfolding and protein-degradation machine.
709 *Biochimica et Biophysica Acta (BBA) - Molecular Cell Research*, AAA ATPases: structure and
710 function 1823, 15–28. <https://doi.org/10.1016/j.bbamcr.2011.06.007>

711 Bassères, A., Tramier, B., 2001. Characterisation of the impact of aqueous industrial waste in mesocosms:
712 biological indicators and pilot streams. *Water Science and Technology* 44, 135–143.
713 <https://doi.org/10.2166/wst.2001.0763>

714 Baussant, T., Bechmann, R.K., Taban, I.C., Larsen, B.K., Tandberg, A.H., Bjørnstad, A., Torgrimsen, S.,
715 Naevdal, A., Øysaed, K.B., Jonsson, G., Sanni, S., 2009. Enzymatic and cellular responses in relation
716 to body burden of PAHs in bivalve molluscs: a case study with chronic levels of North Sea and
717 Barents Sea dispersed oil. *Mar. Pollut. Bull.* 58, 1796–1807.
718 <https://doi.org/10.1016/j.marpolbul.2009.08.007>

719 Benjamini, Y., Hochberg, Y., 1995. Controlling the False Discovery Rate: A Practical and Powerful
720 Approach to Multiple Testing. *Journal of the Royal Statistical Society. Series B (Methodological)* 57,
721 289–300.

722 Bernstein, H.S., Coughlin, S.R., 1998. A Mammalian Homolog of Fission Yeast Cdc5 Regulates G2
723 Progression and Mitotic Entry. *J. Biol. Chem.* 273, 4666–4671. <https://doi.org/10.1074/jbc.273.8.4666>

724 Besse, A., Wu, P., Bruni, F., Donti, T., Graham, B.H., Craigen, W.J., McFarland, R., Moretti, P., Lalani, S.,
725 Scott, K.L., Taylor, R.W., Bonnen, P.E., 2015. The GABA Transaminase, ABAT, Is Essential for
726 Mitochondrial Nucleoside Metabolism. *Cell Metabolism* 21, 417–427.
727 <https://doi.org/10.1016/j.cmet.2015.02.008>

728 Bhattacharyya, S., Klerks, P.L., Nyman, J.A., 2003. Toxicity to freshwater organisms from oils and oil spill
729 chemical treatments in laboratory microcosms. *Environmental Pollution* 122, 205–215.
730 [https://doi.org/10.1016/S0269-7491\(02\)00294-4](https://doi.org/10.1016/S0269-7491(02)00294-4)

731 Bindea, G., Galon, J., Mlecnik, B., 2013. CluePedia Cytoscape plugin: pathway insights using integrated
732 experimental and in silico data. *Bioinformatics* 29, 661–663.
733 <https://doi.org/10.1093/bioinformatics/btt019>

734 Bindea, G., Mlecnik, B., Hackl, H., Charoentong, P., Tosolini, M., Kirilovsky, A., Fridman, W.-H., Pagès,
735 F., Trajanoski, Z., Galon, J., 2009. ClueGO: a Cytoscape plug-in to decipher functionally grouped
736 gene ontology and pathway annotation networks. *Bioinformatics* 25, 1091–1093.
737 <https://doi.org/10.1093/bioinformatics/btp101>

738 Bogenhagen, D.F., Rousseau, D., Burke, S., 2008. The Layered Structure of Human Mitochondrial DNA
739 Nucleoids. *J. Biol. Chem.* 283, 3665–3675. <https://doi.org/10.1074/jbc.M708444200>

740 Boutet, I., Tanguy, A., Moraga, D., 2004. Response of the Pacific oyster *Crassostrea gigas* to hydrocarbon
741 contamination under experimental conditions. *Gene* 329, 147–157.
742 <https://doi.org/10.1016/j.gene.2003.12.027>

743 Boutros, R., Fanayan, S., Shehata, M., Byrne, J.A., 2004. The tumor protein D52 family: many pieces,
744 many puzzles. *Biochemical and Biophysical Research Communications* 325, 1115–1121.
745 <https://doi.org/10.1016/j.bbrc.2004.10.112>

746 Britton, J.C., Morton, B., 1982. A dissection guide, field and laboratory manual for the introduced bivalve
747 *Corbicula fluminea*. *Malacological Review*.

- 748 Brose, N., Petrenko, A.G., Sudhof, T.C., Jahn, R., 1992. Synaptotagmin: a calcium sensor on the synaptic
749 vesicle surface. *Science* 256, 1021–1025. <https://doi.org/10.1126/science.1589771>
- 750 Burlando, B., Berti, E., Viarengo, A., 2006. Effects of seawater pollutants on protein tyrosine
751 phosphorylation in mussel tissues. *Aquatic Toxicology, The Stavanger Workshop Biological Effects of*
752 *Environmental Pollution (BEEP) in marine coastal ecosystem* 78, S79–S85.
753 <https://doi.org/10.1016/j.aquatox.2006.02.020>
- 754 Cailleaud, K., Bassères, A., Gelber, C., Postma, J.F., Schure, A.T.M. ter, Leonards, P.E.G., Redman, A.D.,
755 Whale, G.F., Spence, M.J., Hjort, M., 2019. Investigating predictive tools for refinery effluent hazard
756 assessment using stream mesocosms. *Environmental Toxicology and Chemistry* 38, 650–659.
757 <https://doi.org/10.1002/etc.4338>
- 758 Campos, A., Tedesco, S., Vasconcelos, V., Cristobal, S., 2012. Proteomic research in bivalves: Towards the
759 identification of molecular markers of aquatic pollution. *Journal of Proteomics, Special Issue: Farm*
760 *Animal Proteomics* 75, 4346–4359. <https://doi.org/10.1016/j.jprot.2012.04.027>
- 761 Chantler, P.D., 2006. Chapter 4 Scallop adductor muscles: Structure and function, in: Shumway, S.E.,
762 Parsons, G.J. (Eds.), *Developments in Aquaculture and Fisheries Science, Scallops: Biology, Ecology*
763 *and Aquaculture*. Elsevier, pp. 229–316. [https://doi.org/10.1016/S0167-9309\(06\)80031-1](https://doi.org/10.1016/S0167-9309(06)80031-1)
- 764 Chapman, E.R., 2002. Synaptotagmin: a Ca(2+) sensor that triggers exocytosis? *Nat. Rev. Mol. Cell Biol.* 3,
765 498–508. <https://doi.org/10.1038/nrm855>
- 766 Châtel, A., Talarmin, H., Hamer, B., Schröder, H.C., Müller, W.E.G., Dorange, G., 2011. MAP kinase cell
767 signaling pathway as biomarker of environmental pollution in the sponge *Suberites domuncula*.
768 *Ecotoxicology* 20, 1727–1740. <https://doi.org/10.1007/s10646-011-0706-1>
- 769 Chen, H., Zha, J., Liang, X., Bu, J., Wang, M., Wang, Z., 2013. Sequencing and De Novo Assembly of the
770 Asian Clam (*Corbicula fluminea*) Transcriptome Using the Illumina GAIIx Method. *PLoS One* 8.
771 <https://doi.org/10.1371/journal.pone.0079516>
- 772 Choi, D.W., 1988. Glutamate neurotoxicity and diseases of the nervous system. *Neuron* 1, 623–634.
773 [https://doi.org/10.1016/0896-6273\(88\)90162-6](https://doi.org/10.1016/0896-6273(88)90162-6)
- 774 Cox, G.K., Crossley, D.A., Stieglitz, J.D., Heuer, R.M., Benetti, D.D., Grosell, M., 2017. Oil Exposure
775 Impairs In Situ Cardiac Function in Response to β -Adrenergic Stimulation in Cobia (*Rachycentron*
776 *canadum*). *Environ. Sci. Technol.* 51, 14390–14396. <https://doi.org/10.1021/acs.est.7b03820>
- 777 Danovaro, R., Carugati, L., Berzano, M., Cahill, A.E., Carvalho, S., Chenuil, A., Corinaldesi, C., Cristina,
778 S., David, R., Dell'Anno, A., Dzhenbekova, N., Garcés, E., Gasol, J.M., Goela, P., Féral, J.-P.,
779 Ferrera, I., Forster, R.M., Kurekin, A.A., Rastelli, E., Marinova, V., Miller, P.I., Moncheva, S.,
780 Newton, A., Pearman, J.K., Pitois, S.G., Reñé, A., Rodríguez-Ezpeleta, N., Saggiomo, V., Simis,
781 S.G.H., Stefanova, K., Wilson, C., Lo Martire, M., Greco, S., Cochrane, S.K.J., Mangoni, O., Borja,
782 A., 2016. Implementing and Innovating Marine Monitoring Approaches for Assessing Marine
783 Environmental Status. *Front. Mar. Sci.* 3. <https://doi.org/10.3389/fmars.2016.00213>
- 784 Dongre, S., Sonwane, D.L., 2014. Seasonal Changes In Lipid Content, In The Adductor Muscles Of
785 Cerebralectomied Freshwater Bivalve Mussel *Lamellidens Corrianus*. *IOSR Journal of Pharmacy and*
786 *Biological Sciences* 9, 29–32. <https://doi.org/10.9790/3008-09122932>
- 787 Eganhouse, R.P., Calder, J.A., 1976. The solubility of medium molecular weight aromatic hydrocarbons
788 and the effects of hydrocarbon co-solutes and salinity. *Geochimica et Cosmochimica Acta* 40, 555–
789 561. [https://doi.org/10.1016/0016-7037\(76\)90223-4](https://doi.org/10.1016/0016-7037(76)90223-4)
- 790 Farah, C.S., Reinach, F.C., 1995. The troponin complex and regulation of muscle contraction. *The FASEB*
791 *Journal* 9, 755–767. <https://doi.org/10.1096/fasebj.9.9.7601340>
- 792 Foster, G.D., Baksi, S.M., Means, J.C., 1987. Bioaccumulation of trace organic contaminants from sediment
793 by baltic clams (*Macoma balthica*) and soft-shell clams (*Mya arenaria*). *Environmental Toxicology*
794 *and Chemistry* 6, 969–976. <https://doi.org/10.1002/etc.5620061209>
- 795 Frapiccini, E., Annibaldi, A., Betti, M., Polidori, P., Truzzi, C., Marini, M., 2018. Polycyclic aromatic
796 hydrocarbon (PAH) accumulation in different common sole (*Solea solea*) tissues from the North
797 Adriatic Sea peculiar impacted area. *Marine Pollution Bulletin* 137, 61–68.
798 <https://doi.org/10.1016/j.marpolbul.2018.10.002>
- 799 Gainey, L.F., Walton, J.C., Greenberg, M.J., 2003. Branchial musculature of a venerid clam: pharmacology,
800 distribution, and innervation. *Biol. Bull.* 204, 81–95. <https://doi.org/10.2307/1543498>
- 801 Galler, S., 2008. Molecular basis of the catch state in molluscan smooth muscles: a catchy challenge. *J*
802 *Muscle Res Cell Motil* 29, 73. <https://doi.org/10.1007/s10974-008-9149-6>

803 Goldberg, E.D., Bowen, V.T., Farrington, J.W., Harvey, G., Martin, J.H., Parker, P.L., Risebrough, R.W.,
804 Robertson, W., Schneider, E., Gamble, E., 1978. The Mussel Watch. *Envir. Conserv.* 5, 101–125.
805 <https://doi.org/10.1017/S0376892900005555>

806 Green, J., Trett, M.W. (Eds.), 1989. *The Fate and Effects of Oil in Freshwater*. Springer Netherlands.
807 <https://doi.org/10.1007/978-94-009-1109-3>

808 Hamood, A.W., Marder, E., 2014. Animal-to-Animal Variability in Neuromodulation and Circuit Function.
809 *Cold Spring Harb Symp Quant Biol* 79, 21–28. <https://doi.org/10.1101/sqb.2014.79.024828>

810 Hartmann, J.T., Beggel, S., Auerswald, K., Stoeckle, B.C., Geist, J., 2016. Establishing mussel behavior as
811 a biomarker in ecotoxicology. *Aquatic Toxicology* 170, 279–288.
812 <https://doi.org/10.1016/j.aquatox.2015.06.014>

813 Jiang, M., Li, L., Li, Y., Shen, G., Shen, X., 2017. Oxidative Stress in Shellfish *Sinonovacula constricta*
814 Exposed to the Water Accommodated Fraction of Zero Sulfur Diesel Oil and Pinghu Crude Oil. *Arch.*
815 *Environ. Contam. Toxicol.* 73, 294–300. <https://doi.org/10.1007/s00244-017-0391-z>

816 Johansen, J.L., Esbaugh, A.J., 2019. Oil-induced responses of cardiac and red muscle mitochondria in red
817 drum (*Sciaenops ocellatus*). *Comparative Biochemistry and Physiology Part C: Toxicology &*
818 *Pharmacology* 219, 35–41. <https://doi.org/10.1016/j.cbpc.2019.02.003>

819 Johnson, G.L., Lapadat, R., 2002. Mitogen-activated protein kinase pathways mediated by ERK, JNK, and
820 p38 protein kinases. *Science* 298, 1911–1912. <https://doi.org/10.1126/science.1072682>

821 Käll, L., Canterbury, J.D., Weston, J., Noble, W.S., MacCoss, M.J., 2007. Semi-supervised learning for
822 peptide identification from shotgun proteomics datasets. *Nat. Methods* 4, 923–925.
823 <https://doi.org/10.1038/nmeth1113>

824 Kang, Y., Zhang, X., Dobie, F., Wu, H., Craig, A.M., 2008. Induction of GABAergic Postsynaptic
825 Differentiation by α -Neurexins. *J. Biol. Chem.* 283, 2323–2334.
826 <https://doi.org/10.1074/jbc.M703957200>

827 Kasashima, K., Sumitani, M., Endo, H., 2012. Maintenance of mitochondrial genome distribution by
828 mitochondrial AAA+ protein ClpX. *Experimental Cell Research* 318, 2335–2343.
829 <https://doi.org/10.1016/j.yexcr.2012.07.012>

830 Kerr, A.M., Reisinger, E., Jonas, P., 2008. Differential dependence of phasic transmitter release on
831 synaptotagmin 1 at GABAergic and glutamatergic hippocampal synapses. *Proceedings of the National*
832 *Academy of Sciences* 105, 15581–15586. <https://doi.org/10.1073/pnas.0800621105>

833 Kirby, A.R., Cox, G.K., Nelson, D., Heuer, R.M., Stieglitz, J.D., Benetti, D.D., Grosell, M., Crossley, D.A.,
834 2019. Acute crude oil exposure alters mitochondrial function and ADP affinity in cardiac muscle
835 fibers of young adult Mahi-mahi (*Coryphaena hippurus*). *Comparative Biochemistry and Physiology*
836 *Part C: Toxicology & Pharmacology* 218, 88–95. <https://doi.org/10.1016/j.cbpc.2019.01.004>

837 Kraemer, L.R., Lott, S. 1977. Microscopic anatomy of the visceral mass of Corbicula (Bivalvia;
838 Sphaeriacea). *Bull. Amer. Malacol. Union* 1977: 48-56.

839 Lacroix, C., Le Cuff, N., Receveur, J., Moraga, D., Auffret, M., Guyomarch, J., 2014. Development of an
840 innovative and “green” stir bar sorptive extraction–thermal desorption–gas chromatography–tandem
841 mass spectrometry method for quantification of polycyclic aromatic hydrocarbons in marine biota.
842 *Journal of Chromatography A* 1349, 1–10. <https://doi.org/10.1016/j.chroma.2014.04.094>

843 Lamkanfi, M., Kanneganti, T.-D., 2010. Caspase-7: a protease involved in apoptosis and inflammation. *Int.*
844 *J. Biochem. Cell Biol.* 42, 21–24. <https://doi.org/10.1016/j.biocel.2009.09.013>

845 Lemos, M.F.L., Soares, A.M.V.M., Correia, A.C., Esteves, A.C., 2010. Proteins in ecotoxicology - how,
846 why and why not? *Proteomics* 10, 873–887. <https://doi.org/10.1002/pmic.200900470>

847 Liu, D., Pan, L., Yang, H., Wang, J., 2014. A physiologically based toxicokinetic and toxicodynamic model
848 links the tissue distribution of benzo[a]pyrene and toxic effects in the scallop *Chlamys farreri*.
849 *Environmental Toxicology and Pharmacology* 37, 493–504. <https://doi.org/10.1016/j.etap.2014.01.005>

850 López-Landavery, E.A., Amador-Cano, G., Alejandri, N., Ramirez-Álvarez, N., Montelongo, I., Díaz, F.,
851 Galindo-Sánchez, C.E., 2019. Transcriptomic response and hydrocarbon accumulation in the eastern
852 oyster (*Crassostrea virginica*) exposed to crude oil. *Comparative Biochemistry and Physiology Part C:*
853 *Toxicology & Pharmacology* 225, 108571. <https://doi.org/10.1016/j.cbpc.2019.108571>

854 Lunt, G.G., 1991. GABA and GABA receptors in invertebrates. *Seminars in Neuroscience, GABA and*
855 *Inhibitory Synaptic Transmission in the Brain* 3, 251–258. [https://doi.org/10.1016/1044-5765\(91\)90022-G](https://doi.org/10.1016/1044-5765(91)90022-G)

857 Lunt, S.Y., Fendt, S.-M., 2018. Metabolism – A cornerstone of cancer initiation, progression, immune
858 evasion and treatment response. *Current Opinion in Systems Biology* 8, 67–72.
859 <https://doi.org/10.1016/j.coisb.2017.12.006>

860 Mason, R.P., 1988. Accumulation and depuration of petroleum hydrocarbons by black mussels. 1.
861 Laboratory exposure trials. *South African Journal of Marine Science* 6, 143–153.
862 <https://doi.org/10.2989/025776188784480582>

863 McColl, J.D., Rossiter, R.J., 1951. Lipids of the nervous system of the squid *Loligo pealii*. *J. Exp. Biol.* 28,
864 116–124.

865 Medler, S., Silverman, H., 2001. Muscular alteration of gill geometry in vitro: implications for bivalve
866 pumping processes. *Biol. Bull.* 200, 77–86. <https://doi.org/10.2307/1543087>

867 Miller, M.W., 2019. GABA as a Neurotransmitter in Gastropod Molluscs. *Biol. Bull.* 236, 144–156.
868 <https://doi.org/10.1086/701377>

869 Millman, B.M., 1964. Contraction in the opaque part of the adductor muscle of the oyster (*Crassostrea*
870 *angulata*). *J Physiol* 173, 238–262.

871 Miserazzi, A., Sow, M., Gelber, C., Charifi, M., Ciret, P., Dalens, J.M., Weber, C., Le Floch, S., Lacroix,
872 C., Blanc, P., Massabuau, J.C., 2020. Asiatic clam *Corbicula fluminea* exhibits distinguishable
873 behavioural responses to crude oil under semi-natural multiple stress conditions. *Aquatic Toxicology*
874 219, 105381. <https://doi.org/10.1016/j.aquatox.2019.105381>

875 Mommsen, T.P., 1984. 7 Metabolism of the Fish Gill, in: Hoar, W.S., Randall, D.J. (Eds.), *Fish Physiology*,
876 Gills. Academic Press, pp. 203–238. [https://doi.org/10.1016/S1546-5098\(08\)60186-7](https://doi.org/10.1016/S1546-5098(08)60186-7)

877 Monteiro, L., Moens, T., Lynen, F., Traunspurger, W., 2019. Effects of the water-soluble fraction of a crude
878 oil on freshwater meiofauna and nematode assemblages. *Ecotoxicology and Environmental Safety*
879 176, 186–195. <https://doi.org/10.1016/j.ecoenv.2019.03.083>

880 Mu, R., Wang, Y.-B., Wu, M., Yang, Y., Song, W., Li, T., Zhang, W.-N., Tan, B., Li, A.-L., Wang, N., Xia,
881 Q., Gong, W.-L., Wang, C.-G., Zhou, T., Guo, N., Sang, Z.-H., Li, H.-Y., 2014. Depletion of pre-
882 mRNA splicing factor Cdc5L inhibits mitotic progression and triggers mitotic catastrophe. *Cell Death*
883 *Dis* 5, e1151. <https://doi.org/10.1038/cddis.2014.117>

884 Müller, G. do A.E.S., Lückmann, K.H., Razzera, G., Toledo-Silva, G., Bebianno, M.J., Marques, M.R.F.,
885 Bairy, A.C.D., 2018. Proteomic response of gill microsomes of *Crassostrea brasiliensis* exposed to
886 diesel fuel water-accommodated fraction. *Aquat. Toxicol.* 201, 109–118.
887 <https://doi.org/10.1016/j.aquatox.2018.06.001>

888 Nelson, D., Heuer, R.M., Cox, G.K., Stieglitz, J.D., Hoenig, R., Mager, E.M., Benetti, D.D., Grosell, M.,
889 Crossley, D.A., 2016. Effects of crude oil on in situ cardiac function in young adult mahi-mahi
890 (*Coryphaena hippurus*). *Aquatic Toxicology* 180, 274–281.
891 <https://doi.org/10.1016/j.aquatox.2016.10.012>

892 Nelson, D., Stieglitz, J.D., Cox, G.K., Heuer, R.M., Benetti, D.D., Grosell, M., Crossley, D.A., 2017.
893 Cardio-respiratory function during exercise in the cobia, *Rachycentron canadum*: The impact of crude
894 oil exposure. *Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology* 201,
895 58–65. <https://doi.org/10.1016/j.cbpc.2017.08.006>

896 Nourse, C.R., Mattei, M.-G., Gunning, P., Byrne, J.A., 1998. Cloning of a third member of the D52 gene
897 family indicates alternative coding sequence usage in D52-like transcripts. *Biochimica et Biophysica*
898 *Acta (BBA) - Gene Structure and Expression* 1443, 155–168. [https://doi.org/10.1016/S0167-4781\(98\)00211-5](https://doi.org/10.1016/S0167-4781(98)00211-5)

900 Ogura, T., Wilkinson, A.J., 2001. AAA+ superfamily ATPases: common structure--diverse function. *Genes*
901 *Cells* 6, 575–597.

902 Pruell, R.J., Lake, J.L., Davis, W.R., Quinn, J.G., 1986. Uptake and depuration of organic contaminants by
903 blue mussels (*Mytilus edulis*) exposed to environmentally contaminated sediment. *Marine Biology* 91,
904 497–507. <https://doi.org/10.1007/BF00392601>

905 Pyle, G., Ford, A.T., 2017. Behaviour revised: Contaminant effects on aquatic animal behaviour. *Aquatic*
906 *Toxicology* 182, 226–228. <https://doi.org/10.1016/j.aquatox.2016.11.008>

907 R Core Team, 2016. R: A language and environment for statistical computing. R Foundation for Statistical
908 Computing, Vienna, Austria. URL <https://www.R-project.org/>.

909 Ramachandran, S.D., Swezey, M.J., Hodson, P.V., Boudreau, M., Courtenay, S.C., Lee, K., King, T.,
910 Dixon, J.A., 2006. Influence of salinity and fish species on PAH uptake from dispersed crude oil.
911 *Marine Pollution Bulletin* 52, 1182–1189. <https://doi.org/10.1016/j.marpolbul.2006.02.009>

912 Reissner, C., Runkel, F., Missler, M., 2013. Neurexins. *Genome Biology* 14, 213.
913 <https://doi.org/10.1186/gb-2013-14-9-213>

914 Rivière, G., Klopp, C., Ibouniyamine, N., Huvet, A., Boudry, P., Favrel, P., 2015. GigaTON: an extensive
915 publicly searchable database providing a new reference transcriptome in the pacific oyster *Crassostrea*
916 *gigas*. *BMC Bioinformatics* 16, 401. <https://doi.org/10.1186/s12859-015-0833-4>

917 Roberts, R.J., Agius, C., Saliba, C., Bossier, P., Sung, Y.Y., 2010. Heat shock proteins (chaperones) in fish
918 and shellfish and their potential role in relation to fish health: a review. *J. Fish Dis.* 33, 789–801.
919 <https://doi.org/10.1111/j.1365-2761.2010.01183.x>

920 Rossi, S.S., Thomas, W.H., 1981. Solubility behavior of three aromatic hydrocarbons in distilled water and
921 natural seawater. *Environ. Sci. Technol.* 15, 715–716. <https://doi.org/10.1021/es00088a013>

922 Salazar-Coria, L., Rocha-Gómez, M.A., Matadamas-Martínez, F., Yépez-Mulia, L., Vega-López, A., 2019.
923 Proteomic analysis of oxidized proteins in the brain and liver of the Nile tilapia (*Oreochromis*
924 *niloticus*) exposed to a water-accommodated fraction of Maya crude oil. *Ecotoxicol. Environ. Saf.*
925 171, 609–620. <https://doi.org/10.1016/j.ecoenv.2019.01.033>

926 Shannon, P., Markiel, A., Ozier, O., Baliga, N.S., Wang, J.T., Ramage, D., Amin, N., Schwikowski, B.,
927 Ideker, T., 2003. Cytoscape: a software environment for integrated models of biomolecular interaction
928 networks. *Genome Res.* 13, 2498–2504. <https://doi.org/10.1101/gr.1239303>

929 Shen, Y., Wen, Q., Liu, H., Zhong, C., Qin, Y., Harris, G., Kawano, T., Wu, M., Xu, T., Samuel, A.D.,
930 Zhang, Y., 2016. An extrasynaptic GABAergic signal modulates a pattern of forward movement in
931 *Caenorhabditis elegans*. *eLife* 5, e14197. <https://doi.org/10.7554/eLife.14197>

932 Shrivastava, A., Gupta, V.B., 2011. Methods for the determination of limit of detection and limit of
933 quantitation of the analytical methods. *Chronicles of Young Scientists* 2, 21.
934 <https://doi.org/10.4103/2229-5186.79345>

935 Shukla, P., Gopalani, M., Ramteke, D.S., Wate, S.R., 2007. Influence of Salinity on PAH Uptake from
936 Water Soluble Fraction of Crude Oil in *Tilapia mossambica*. *Bull Environ Contam Toxicol* 79, 601–
937 605. <https://doi.org/10.1007/s00128-007-9272-x>

938 Sies, H., 2017. Hydrogen peroxide as a central redox signaling molecule in physiological oxidative stress:
939 Oxidative eustress. *Redox Biol* 11, 613–619. <https://doi.org/10.1016/j.redox.2016.12.035>

940 Stegeman, J.J., Teal, J.M., 1973. Accumulation, release and retention of petroleum hydrocarbons by the
941 oyster *Crassostrea virginica*. *Marine Biology* 22, 37–44. <https://doi.org/10.1007/BF00388908>

942 Sun, X., Liu, Z., Wu, B., Zhou, L., Wang, Q., Wu, W., Yang, A., 2018. Differences between fast and slow
943 muscles in scallops revealed through proteomics and transcriptomics. *BMC Genomics* 19, 377.
944 <https://doi.org/10.1186/s12864-018-4770-2>

945 Sysoeva, T.A., 2017. Assessing heterogeneity in oligomeric AAA+ machines. *Cell. Mol. Life Sci.* 74,
946 1001–1018. <https://doi.org/10.1007/s00018-016-2374-z>

947 Tanaka, H., Suzuki, H., Ohtsuki, I., Ojima, T., 2008. Structure-function relationships of molluscan troponin
948 T revealed by limited proteolysis. *Biochim. Biophys. Acta* 1784, 1037–1042.
949 <https://doi.org/10.1016/j.bbapap.2008.04.001>

950 Tomanek, L., 2015. Proteomic responses to environmentally induced oxidative stress. *Journal of*
951 *Experimental Biology* 218, 1867–1879. <https://doi.org/10.1242/jeb.116475>

952 Tran, D., Ciret, P., Ciutat, A., Durrieu, G., Massabuau, J.-C., 2003. Estimation of potential and limits of
953 bivalve closure response to detect contaminants: Application to cadmium. *Environmental Toxicology*
954 *and Chemistry* 22, 914–920. <https://doi.org/10.1002/etc.5620220432>

955 Tyanova, S., Cox, J., 2018. Perseus: A Bioinformatics Platform for Integrative Analysis of Proteomics Data
956 in Cancer Research, in: von Stechow, L. (Ed.), *Cancer Systems Biology: Methods and Protocols*,
957 *Methods in Molecular Biology*. Springer New York, New York, NY, pp. 133–148.
958 https://doi.org/10.1007/978-1-4939-7493-1_7

959 Tyanova, S., Temu, T., Sinitcyn, P., Carlson, A., Hein, M.Y., Geiger, T., Mann, M., Cox, J., 2016. The
960 Perseus computational platform for comprehensive analysis of (prote)omics data. *Nature Methods* 13,
961 731–740. <https://doi.org/10.1038/nmeth.3901>

962 Van Aggelen, G., Ankley, G.T., Baldwin, W.S., Bearden, D.W., Benson, W.H., Chipman, J.K., Collette,
963 T.W., Craft, J.A., Denslow, N.D., Embry, M.R., Falciani, F., George, S.G., Helbing, C.C., Hoekstra,
964 P.F., Iguchi, T., Kagami, Y., Katsiadaki, I., Kille, P., Liu, L., Lord, P.G., McIntyre, T., O'Neill, A.,
965 Osachoff, H., Perkins, E.J., Santos, E.M., Skirrow, R.C., Snape, J.R., Tyler, C.R., Versteeg, D., Viant,

966 M.R., Volz, D.C., Williams, T.D., Yu, L., 2010. Integrating Omic Technologies into Aquatic
967 Ecological Risk Assessment and Environmental Monitoring: Hurdles, Achievements, and Future
968 Outlook. *Environmental Health Perspectives* 118, 1–5. <https://doi.org/10.1289/ehp.0900985>
969 van Haren, R.J.F., Schepers, H.E., Kooijman, S.A.L.M., 1994. Dynamic energy budgets affect kinetics of
970 xenobiotics in the marine mussel *Mytilus edulis*. *Chemosphere* 29, 163–189.
971 [https://doi.org/10.1016/0045-6535\(94\)90099-X](https://doi.org/10.1016/0045-6535(94)90099-X)
972 Vandermeulen, J.H., Ross, C.W., 1995. Oil spill response in freshwater: Assessment of the impact of
973 cleanup as a management tool. *Journal of Environmental Management* 44, 297–308.
974 [https://doi.org/10.1016/S0301-4797\(95\)90338-0](https://doi.org/10.1016/S0301-4797(95)90338-0)
975 Walker, M.C., van der Donk, W.A., 2016. The many roles of glutamate in metabolism. *J. Ind. Microbiol.*
976 *Biotechnol.* 43, 419–430. <https://doi.org/10.1007/s10295-015-1665-y>
977 Walls, A.B., Waagepetersen, H.S., Bak, L.K., Schousboe, A., Sonnewald, U., 2015. The glutamine-
978 glutamate/GABA cycle: function, regional differences in glutamate and GABA production and effects
979 of interference with GABA metabolism. *Neurochem. Res.* 40, 402–409.
980 <https://doi.org/10.1007/s11064-014-1473-1>
981 Webster, L., Russell, M., Packer, G., Moffat, C.F., 2006. Long term monitoring of polycyclic aromatic
982 hydrocarbons (PAHs) in blue mussels (*Mytilus edulis*) from a remote Scottish location. *Polycyclic*
983 *Aromatic Compounds* 26, 283–298. <https://doi.org/10.1080/10406630600904109>
984 Won, E.-J., Kim, R.-O., Kang, H.-M., Kim, H.-S., Hwang, D.-S., Han, J., Lee, Y.H., Hwang, U.-K., Zhou,
985 B., Lee, S.-J., Lee, J.-S., 2016. Adverse Effects, Expression of the Bk-CYP3045C1 Gene, and
986 Activation of the ERK Signaling Pathway in the Water Accommodated Fraction-Exposed Rotifer.
987 *Environ. Sci. Technol.* 50, 6025–6035. <https://doi.org/10.1021/acs.est.6b01306>
988 Wong, B.B.M., Candolin, U., 2015. Behavioral responses to changing environments. *Behavioral Ecology*
989 26, 665–673. <https://doi.org/10.1093/beheco/aru183>
990 Wu, H., Wang, W.-X., 2010. NMR-based metabolomic studies on the toxicological effects of cadmium and
991 copper on green mussels *Perna viridis*. *Aquat. Toxicol.* 100, 339–345.
992 <https://doi.org/10.1016/j.aquatox.2010.08.005>
993 Wu, W.-J., Qin, N., He, W., He, Q.-S., Ouyang, H.-L., Xu, F.-L., 2012. Levels, Distribution, and Health
994 Risks of Polycyclic Aromatic Hydrocarbons in Four Freshwater Edible Fish Species from the Beijing
995 Market. *The Scientific World Journal* 2012, 1–12. <https://doi.org/10.1100/2012/156378>
996 Wu, Y., Brosh, R.M., 2010. Distinct roles of RECQ1 in the maintenance of genomic stability. *DNA Repair*
997 (Amst.) 9, 315–324. <https://doi.org/10.1016/j.dnarep.2009.12.010>
998 Xu, F.-L., Wu, W.-J., Wang, J.-J., Qin, N., Wang, Y., He, Q.-S., He, W., Tao, S., 2011. Residual levels and
999 health risk of polycyclic aromatic hydrocarbons in freshwater fishes from Lake Small Bai-Yang-Dian,
1000 Northern China. *Ecological Modelling* 222, 275–286. <https://doi.org/10.1016/j.ecolmodel.2010.10.001>
1001 Yamada, A., Yoshio, M., Oiwa, K., 2013. Myosin Mg-ATPase of molluscan muscles is slightly activated by
1002 F-actin under catch state in vitro. *Journal of Muscle Research and Cell Motility* 34, 115–123.
1003 <https://doi.org/10.1007/s10974-013-9339-8>
1004 Zhang, Guofan, Fang, X., Guo, X., Li, L., Luo, R., Xu, F., Yang, P., Zhang, L., Wang, X., Qi, H., Xiong, Z.,
1005 Que, H., Xie, Y., Holland, P.W.H., Paps, J., Zhu, Y., Wu, F., Chen, Y., Wang, Jiafeng, Peng, C.,
1006 Meng, J., Yang, L., Liu, J., Wen, B., Zhang, N., Huang, Z., Zhu, Q., Feng, Y., Mount, A., Hedgecock,
1007 D., Xu, Z., Liu, Y., Domazet-Lošo, T., Du, Y., Sun, X., Zhang, Shoudu, Liu, B., Cheng, P., Jiang, X.,
1008 Li, J., Fan, D., Wang, W., Fu, W., Wang, T., Wang, B., Zhang, J., Peng, Z., Li, Yingxiang, Li, Na,
1009 Wang, Jinpeng, Chen, M., He, Y., Tan, F., Song, X., Zheng, Q., Huang, R., Yang, Hailong, Du, X.,
1010 Chen, L., Yang, M., Gaffney, P.M., Wang, S., Luo, L., She, Z., Ming, Y., Huang, W., Zhang, Shu,
1011 Huang, B., Zhang, Y., Qu, T., Ni, P., Miao, G., Wang, Junyi, Wang, Q., Steinberg, C.E.W., Wang, H.,
1012 Li, Ning, Qian, L., Zhang, Guojie, Li, Yingrui, Yang, Huanming, Liu, X., Wang, Jian, Yin, Y., Wang,
1013 Jun, 2012. The oyster genome reveals stress adaptation and complexity of shell formation. *Nature* 490,
1014 49–54. <https://doi.org/10.1038/nature11413>
1015 Zhang, Yang, Sun, J., Mu, H., Li, J., Zhang, Yuehuan, Xu, F., Xiang, Z., Qian, P.-Y., Qiu, J.-W., Yu, Z.,
1016 2015. Proteomic Basis of Stress Responses in the Gills of the Pacific Oyster *Crassostrea gigas*. *J.*
1017 *Proteome Res.* 14, 304–317. <https://doi.org/10.1021/pr500940s>
Zot, A.S., Potter, J.D., 1987. Structural aspects of troponin-tropomyosin regulation of skeletal muscle
contraction. *Annu Rev Biophys Chem* 16, 535–559.
<https://doi.org/10.1146/annurev.bb.16.060187.002535>