

1 **Title:** First assessment of Atlantic open ocean *Sargassum* spp. metal and metalloid
2 concentrations

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18 **Abstract (150 words)**

19 Over the last decade, increasing proliferations of Atlantic *Sargassum* populations have
20 led to massive beaching with disastrous environmental consequences. This study is a
21 preliminary assessment of open ocean *Sargassum* spp. element concentration to assess their
22 potential contribution on coastal ecosystems. *Sargassum* spp. samples from seven sites,
23 collected along a transect from the center of the Atlantic Ocean to near the coast of
24 Martinique (French West Indies), were analyzed to determine their potential metal and
25 metalloid contamination. Mean element concentrations from the *Sargassum* spp. samples
26 were ranked in the following descending order: As > Fe > Mn > Al > Zn > V > Ni > Cu > Cr
27 > Cd > Hg. Element concentrations are relatively low compared to previous results of
28 beached *Sargassum* spp. except for As that need to be carefully considered before reusing
29 *Sargassum* spp.

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32 **Keywords (6max):**

33 *Sargassum*; metal; metalloid; contamination; Atlantic

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37 **I. Introduction**

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39 The brown algae *Sargassum* is one of the most diverse marine macro-algae genera
40 with 362 taxonomically accepted species with only two pelagic species: *Sargassum fluitans*
41 and *natans*. These two planktonic species form dense population rafts on the ocean surface,
42 mainly of the Atlantic, and play a crucial role in the wider ecosystem. *Sargassum* spp. are
43 home to a wide range of species (many of which are endemic), provide nurseries and cover
44 habitat for species going from invertebrates to fishes (including commercially important fish
45 species), and endangered turtles (Casazza and Ross, 2008; Witherington et al., 2012).

46 The presence of free floating *Sargassum* in the Northern Atlantic Ocean is not a new
47 finding; Christopher Columbus reported seeing *Sargassum* spp. as far back as the 15th
48 century. Since 2011, however, proliferation of *Sargassum* spp. populations have been
49 observed in Africa (Oyesiku and Egunyomi, 2014; Addico and Atta deGraft-Johnson, 2016)
50 and in the Wider Caribbean Region (WCR), including locations where they were so far absent
51 or extremely rare (Hu et al., 2016; Sissini et al., 2017). Massive *Sargassum* spp. landings (i.e.
52 arrival of massive amounts of *Sargassum* on beaches) in the Caribbean area, such as the 2011
53 and the even stronger 2014-2015 events, lead to disastrous consequences with major impacts
54 on public health (Anderson, 2007), fisheries (Solarin et al., 2014), coastal ecosystems
55 (Rodríguez-Martínez et al., 2019), and tourism (Louime, et al., 2017).

56 Despite their harmful consequences, few studies have been carried out on open ocean
57 pelagic *Sargassum* spp. and especially, to our knowledge, none have been done on their
58 potential role as a carrier of metal pollutants. Brown algae, such as *Sargassum* spp.,
59 accumulate heavy metals from the surrounding environment and the concentration in their cell
60 walls may be 20,000 to 40,000 times higher than in the surrounding water (Sudharsan et al.,
61 2012; Sadeghi et al., 2014). The biosorption capability of *Sargassum* spp., or more generally
62 of brown seaweed, have made them flawless biological indicators of heavy metal pollutions
63 (Haug et al., 1974; Butler et al., 1983; Philips, 1990; Khristoforova and Kozhenkova, 2002;
64 Chernova and Sergeeva, 2008; Thangaradjou et al., 2010). Heavy metal concentration in
65 brown algae is proportional to the quantity of bioavailable forms of metals in sea water during
66 the period of algal vegetation (Karthick et al., 2013). Floating *Sargassum* spp. landing in
67 Caribbean coasts are particularly contaminated by As (Devault et al., 2021). *Sargassum* spp.
68 with high levels of metal and metalloid contamination might impact the organisms that feed

69 on them and could also result in the transport of contamination from the open ocean to coastal
70 sites *via* beaching.

71 Metals and metalloids are naturally occurring elements found throughout the earth's
72 crust. Their oceanic distributions are not homogenous, there is areas of higher metal contents
73 such as region under volcanic activities (Kamenev et al., 2004) and upwelling (Bruland, 1980;
74 Yeats and Campeell, 1983). Anthropogenic activities have however released a large number
75 of heavy metals in the environment leading to large contaminations in coastal and marine
76 environments. Metal pollution can dramatically impact human health, aquatic organisms, and
77 natural ecosystems because of their toxicity, persistence, and bioaccumulation characteristics
78 (DeForest et al., 2007; Karthick et al., 2012). As *Sargassum* spp. continues to be beached in
79 large quantities, we must understand their impacts and how to safely reuse them. Various
80 sargassum valorization exists such as fertilizer for agricultural crops, bioenergy or beauty care
81 products (Chávez et al., 2020). *Sargassum* spp. that reach coastal areas and are finally
82 beached may introduce metal and metalloid contamination; it is consequently crucial to
83 determine their toxicity before their valorization. Depending on their origins and their
84 journey, open ocean free floating *Sargassum* spp. might present different level of metal and
85 metalloid contamination. Therefore, this present study tackles the issue of estimating metal
86 and metalloid concentrations in open ocean pelagic *Sargassum* spp. before their arrival in the
87 Caribbean region. To accomplish these measurements, pelagic *Sargassum* spp. were collected
88 in various locations along a transect from the center of the Atlantic Ocean to near the coast of
89 Martinique (French West Indies). We then make comparisons to previously reported values
90 on metal concentrations on both benthic and planktonic *Sargassum* spp. as well as to open
91 ocean metal concentrations. The ultimate goal of this study is to present a preliminary
92 assessment of open ocean pelagic *Sargassum* spp. metal and metalloid contamination before
93 their arrival in the Caribbean Sea.

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96 **II. Material and methods**

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98 **I.I. Samples collection and preparation**

99 In November 2018, as part of a science participative initiative, a volunteer sailor
100 collected pelagic *Sargassum* spp. samples during a transatlantic sailing cruise. Every time the
101 boat encountered *Sargassum* spp. rafts, a sample of *Sargassum* spp. was collected by hand
102 and the geolocation of the ship position was recorded via Global Positioning System (GPS);

103 seven sites were sampled following this method (Figure 1 and Table S1). Once collected,
104 *Sargassum* spp. were rinsed with seawater, dried on the ship hull, and stored in plastic bags to
105 avoid potential metal and metalloid contamination from other sources.

106 In the laboratory, samples were rinsed five times with deionized water and oven-dried
107 at 40°C for 48 hours. There are three main morphotypes of pelagic *Sargassum*: *Sargassum*
108 *fluitans* III and *Sargassum natans* I and VIII. Visual identification of the species was
109 established using criteria described in Oyesiku and Egunyomi (2014) and in Fernandez et al.
110 (2017). *Sargassum* spp. in our samples were a mix of the three morphotypes. We did not
111 have enough material, neither enough sites, to be able to discriminate data by morphotype.
112 Our samples are therefore a mixture of the three morphotypes and no distinction in regard of
113 the morphotype will therefore be done and for the remaining of the manuscript, we will only
114 refer as *Sargassum* spp.

115

116 **I.2. Element analyses**

117 *Sargassum* spp. samples were ground using an agate mortar into a finely homogenized
118 powder following previously published protocols (e.g. Zou et al., 2015; Kaviarasan et al.,
119 2018; Pan et al., 2018). Powdered samples (100 mg) were weighed and digested with 3 mL of
120 pure nitric acid (67 %). Mineralization was done by heating up the samples at 100 °C for three
121 hours (hot block CAL 3300, Environmental Express, USA), in closed mineralization tubes.
122 After the mineralization, 15 mL of ultra-pure water (Milli-Q, Bedford, MA, USA) was added
123 to each sample. For each site, three aliquots were analyzed except for sites 2 and 4 for which
124 only two aliquots were performed because of the small quantity of *Sargassum* spp. collected
125 (less than 300 mg).

126 A series of 14 elements were analyzed simultaneously by Inductively Coupled Plasma
127 Optical Emission Spectrometer (700 Series ICP-OES, Agilent): Silver (Ag), Aluminum (Al),
128 Arsenic (As), Cadmium (Cd), Cobalt (Co), Chrome (Cr), Copper (Cu), Iron (Fe), Manganese
129 (Mn), Nickel (Ni), Lead (Pb), Selenium (Se), Vanadium (V), and Zinc (Zn). Certified
130 reference materials DOLT-5 (dogfish (*Squalus acanthias*) liver), TORT-3 (Lobster
131 Hepatopancreas), IAEA-413 (Algae) were analyzed using the same methodology as the
132 *Sargassum* spp. samples; their recovery rates vary between 84.58 and 107.59 % (Table S2).
133 Element concentrations in *Sargassum* spp. samples are expressed in $\mu\text{g g}^{-1}$ dry weight (dw).
134 For values below the instrument detection limit, theoretical minimum concentration values are
135 calculated (the detection limit of the instrument (in $\mu\text{g.g}^{-1}$) multiplied by the volume of the
136 sample (in L) divided by the sample *Sargassum* weight (in g)).

137 To determine Hg concentrations in *Sargassum* spp. samples, aliquots of ~15 mg of
138 powdered *Sargassum* spp. were analyzed by flameless atomic absorption spectrometry (AMA
139 254, SYMALAB, France). The validity of the analytical method was verified against a
140 biologic reference material IAEA-407 (fish tissue). *Sargassum* spp. Hg concentration and
141 recovery rates of the reference material are presented in Table S2. There are no Hg values
142 from Site # 2 because there was not enough sample left to perform the analysis.

143

144 **I.3. Statistical analyses**

145 Differences in mean element concentrations were tested *via* a student's t-test after a
146 Box-Cox transformation (Peltier et al., 1998). The normality of the variance and
147 homogeneity were validated by a Shapiro's and a Levene's test, respectively. For data that
148 yielded neither normal variance nor homogeneity, the difference of the mean significance was
149 determined using a Mann-Whitney u-test. Principal component analyses (PCA) were obtained
150 using the R Cran software (FactoMineR and factoextra packages). The Kaiser criterion was
151 used to select which dimension could be used for interpretation. Pearson product moment
152 correlation coefficients (referred to as correlation) were used to determine the significance of
153 relationships between elements and the three PCA dimensions as well as the relationship
154 between the different *Sargassum* spp. element concentrations. If not specified, the
155 significance was calculated at the 95 % confidence level. For box plot, clusters with the same
156 letter code are not significantly different at the 95% confidence level (t-test or U-test).
157 Horizontal black lines within the boxes mark the median.

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159

160 **III. Results**

161

162 The element concentrations (and respective standard errors) measured from
163 *Sargassum* spp. collected at the seven sites sampled in this study are presented in Table S2.
164 The three aliquots for Sites 1, 3, 5, 6, and 7 and the two aliquots for sites 2 and 4 yield good
165 reproducibility between replicated with values of relative standard deviation being on average
166 less than 10 % (values not presented here). The quantity of *Sargassum* spp. collected did not
167 permit analysis of more than three aliquots per site but the established reproducibility between
168 them indicates that site-specific *Sargassum* spp. element concentrations are homogeneous and
169 therefore representative of the element concentrations of their *Sargassum* spp. raft source. Out
170 of the 14 elements analyzed, 11 were above detection limits: Al, As, Cd, Cr, Cu, Fe, Mn, Ni,

171 V, Zn, and Hg (Table S2). *Sargassum* spp. Ag, Co, Pb, and Se concentrations are below the
172 detection limits; mean theoretical minimum concentration values and 1σ standard deviation
173 are: 0.0189 ± 0.0021 for Ag, 0.0144 ± 0.0016 for Co, 0.0931 ± 0.0102 for Pb, and $0.1054 \pm$
174 0.0116 for Se. Mean elemental concentrations can be sequenced in the following descending
175 order: As > Fe > Mn > Al > Zn > V > Ni > Cu > Cr > Cd > Hg.

176 To be able to elucidate the role of spatial variability (*i.e.* location) on the concentration
177 of each element, Principal Component Analyses were performed (Figure 2 A et B). The first
178 three dimensions are conserved for interpretation as they passed the Kaiser criterion (the
179 cumulative variance of the first three dimensions explain 90% of the total variance).
180 Correlations between the element concentrations and the three dimensions are presented in
181 Figure 2 C. The dimension 1 explains about 50 % of the total variance (Figure 2.A.).
182 Dimension 1 permits discrimination between Site 1 from Sites 5 and 7. Site 1 is influenced by
183 (and therefore more enriched in) Cr, Fe, As, V, and Al; Sites 5 and 7 are, however, influenced
184 by Cd and Cu. Dimension 2 (Figure 2 A) explains about 22 % of the total variance. Three
185 variables (Mn, Zn, Cu) permit discrimination between Site 6 from Sites 2 and 4; *Sargassum*
186 spp. element concentrations from Site 1 present higher Cu, Mn, and Zn values than Sites 2
187 and 4. Dimension 3 (Figure 2 B) explains about 15 % of the total variance. The variable Ni,
188 which is only associated with this dimension, permits to oppose Sites 2 and 3 from Site 7 as
189 the site largely affected by Ni.

190 To visualize the results generated from this study site-specific elemental
191 concentrations, site-specific box plots of *Sargassum* spp. element concentrations were created
192 (Figure 3). For sites for which it was possible to analyze three aliquots (Sites 1, 3, 5, 6, and 7)
193 significant differences between sites are represented by letters. *Sargassum* spp. from the seven
194 sites present different levels of element concentrations and the element repartition is not
195 homogenous between sites. Four different accumulation profiles are observed: i) a significant
196 (at the 95 % significance level; Table S4) decreasing concentration gradient from Site 1 to
197 Site 7 for Al, As, Fe, V, Hg and into a lesser extend Cr, ii) an asymmetric u-shaped profile for
198 Ni with Site 5 presenting the lowest concentration and Site 7 the highest (1.5-times higher
199 than Site 1), each site presents significantly different values, iii) a two sided profile for Cu
200 with low concentrations for Sites 1 to 4 and significantly higher concentrations for Sites 5 to
201 7, and iv) a random profile for Cd, Mn, and Zn, with Mn and Zn profiles significantly
202 correlated at the 95% significance level (Table S4).

203 To put these results into perspective, we compared them to a review of previously
204 published *Sargassum* spp. metal(loid) concentrations. First, we compare our open ocean

205 pelagic *Sargassum* spp. element concentrations to concentrations from coastal benthic
206 *Sargassum* spp, coming from a large review of already published data (Table S3 and Figure
207 3). When looking at the range of variability, our open ocean pelagic *Sargassum* spp. Cd, Cu,
208 Fe, Ni, and Hg values are significantly lower (at the 95 % significance level) while Cr, Mn
209 and Zn values are on the lower range of coastal benthic *Sargassum* spp literature's values.
210 When examining the median element concentrations, the previously published coastal benthic
211 *Sargassum* spp. studies yielded values that are higher than those generated from our open
212 ocean pelagic *Sargassum* spp. samples, ranging from 1.2-times for Zn to 10-times for Fe.
213 Secondly, we looked at three recently published studies that have analyzed coastal planktonic
214 *Sargassum* spp. (Table 1; values not plotted on Figure 3); one from the Dominican Republic's
215 beaches (Fernández et al., 2017), one from the coast of Ghana (Addico and Atta deGraft-
216 Johnson, 2016), and one from the Mexican Caribbean coast (Rodríguez-Martínez et al.,
217 2020). Compared to our open ocean pelagic *Sargassum* spp. element concentrations,
218 planktonic *Sargassum* spp. from the coast of Ghana (Addico and Atta deGraft-Johnson, 2016)
219 yielded concentrations significantly higher (95 % significance level) for Cd, Cu, Fe, Zn, and
220 Hg and significantly lower for As. *Sargassum* spp. from the Dominican Republic (Fernández
221 et al., 2017; Table 1), also yielded concentration values significantly higher (95 %
222 significance level) for Al, Cr, and Hg, in the same range for Cu, Fe, Mn, Ni, V, and Zn and
223 significantly lower for As and Cd. Finally, *Sargassum* spp. from the Mexican Caribbean coast
224 (Rodríguez-Martínez et al., 2020; Table 1) yielded concentration values significantly higher
225 (95 % significance level) for Al, in the same range for As and Cu, and significantly lower for
226 Fe, Mn, V, and Hg.

227

228

229 **IV. Discussion**

230 Mean elemental concentrations are sequenced in the following descending order: As >
231 Fe > Mn > Al > Zn > V > Ni > Cu > Cr > Cd > Hg. Levels of essential elements (de Boer et
232 al., 1986; Allan, 1997; Rodrigues Silva et al., 2009; Tamilselvan et al., 2012; Rehder, 2015)
233 like Fe, Mn, Zn, V, and Cu are on average higher than levels of non-essential elements
234 (Rodrigues Silva et al., 2009; Yusuf et al., 2011) like Ni and Cr. The non-essential and mainly
235 toxic elements such as Hg and Cd (Allan, 1997; Tamilselvan et al., 2012; Costa et al., 2017),
236 despite not being highly concentrated, are present in *Sargassum* spp. Both Al and As, in spite
237 of being non-essential elements (Rybak et al., 2017; Al Mamum et al., 2019; Ameri et al.,

238 2020), are present in higher concentrations than most of the essential ones, a distinctive
239 feature already observed in other studies (see Maret, 2016 for a review).

240 Open ocean pelagic *Sargassum* spp. from this study were collected in November 2018;
241 during that year, the month with the highest coverage (i.e. surface area) of oceanic *Sargassum*
242 spp. near the Mexican Caribbean coastline was recorded in September with 22,900 ha
243 (Chávez et al., 2020). Despite this clear *Sargassum* spp. peak, surface area varied
244 considerably without any clear pattern throughout the rest of the year (Chávez et al., 2020).
245 This variability could explain the relatively low quantity of *Sargassum* spp. rafts encountered
246 and collected during our November cruise compared to the other months.

247 Results from this study also reveal heterogeneous element repartition in regard to
248 spatial variation. The four distinct accumulation profiles indicate the absence of a clear
249 longitudinal concentration gradient between sites. Element concentrations in *Sargassum* spp.
250 reflect the presence of these elements in ocean water in dissolved forms. The relative
251 concentration of one element versus another as recorded by *Sargassum* spp., however, is not
252 representative of the ratio in seawater. This is because *Sargassum* spp. does not
253 bioaccumulate each element in an identical manner (e.g. Abirhire and Kadiri, 2011; Sadeghi
254 et al., 2014; Chen et al., 2018). Various factors can influence the presence, distribution, and
255 variability of elements in open ocean. Among the natural factors, a potential source of
256 metal and metalloid concentration heterogeneity could be the proximity to upwelling as they
257 are the principal source of surface nutrients and therefore trace metals from the deep ocean
258 (Valdés et al., 2008). Higher concentration of dissolved Zn, Ni, Co and Cd were found in
259 upwelling surface and sub-surface waters (Bruland, 1980; Yeats and Campeell, 1983; Valdés
260 et al., 2008; Kavun and Podgurskaya, 2009; Ahlgren et al. (2014)). *Sargassum* sp. from this
261 study do not present specific enrichment in those metal, leading to believe that they might not
262 originated from nor crossed upwelling areas.

263 The variability of metal and metalloid concentrations in seawater may also be
264 impacted by seasonality and time of day (Philips, 1977). Algae chemical compositions are
265 sensitive to the soluble-trace metal content of their ambient surroundings. However, they do
266 not represent the total metal loads in seawater, as they cannot incorporate metals associated
267 with organic or inorganic particulate matter (Philips, 1977). Moreover, variability in short-
268 term oceanic metal concentrations cannot be responsible for inter-site variability in
269 *Sargassum* spp. since the algae integrate contaminants over their entire life span.

270 The program GEOTRACES (<https://www.geotraces.org/>) surveys critical regions of
271 the world's oceans measuring trace elements and their isotopes that are known indicators of

272 important biogeochemical and physical processes. Seawater metal concentrations data from
273 the GEOTRACES dissolved elements database, separated into 5 boxes to represent the
274 position of this present study's sites, are presented in Table 2. The significance of the
275 difference between the different GEOTRACES boxes cannot be established as the number of
276 values per box are less than three. However, a comparison between GEOTRACES and our
277 *Sargassum* spp. data do not reveal a similar pattern. For example, high *Sargassum* spp. values
278 of Fe and Al that discriminate Site 1 from Sites 5, 6, and 7 do not present a similar pattern in
279 GEOTRACES data; boxes representatives of Sites 5, 6, and 7 present similar or higher values
280 of Fe and Al compared to the box representative of Site 1. The absence of a perfect match
281 between *Sargassum* spp. and GEOTRACES metal concentrations is not surprising as
282 GEOTRACES data are from samples taken at a discrete moment in time and space while the
283 *Sargassum* spp. data, as said previously, integrate metal concentrations over their entire life
284 span.

285 A potential external source of metals and metalloids into the ocean is fluvial inputs *via*
286 river runoff. Various studies have described a decrease in metal concentrations along transects
287 from coastal waters to the open ocean; only a fraction of metals leave the coastal zone and are
288 transported to the open ocean by advection-diffusion processes (Schaule and Patterson, 1981;
289 Symes and Kester, 1985; Landing and Bruland, 1987; Martin and Gordon, 1988). Metals can
290 also enter the ocean by atmospheric inputs; it is well established that the Pb found in the
291 Atlantic Ocean results from the atmospheric input dispersed from the North American
292 continent *via* the westerly winds (Mart et al., 1982). Atmospheric deposition of mercury from
293 continental origin contributes significantly to the variability of surface ocean mixed layer
294 (Zhang et al., 2016). If *Sargassum* spp. element concentrations were influenced by fluvial
295 and/or atmospheric inputs, sites closer to the coast should present higher elemental
296 concentrations. Site 1, the farthest away from the coast, however, yields the highest
297 concentrations in Al, As, Cr, Fe, V, and Hg, which dismantles this hypothesis. This
298 hypothesis is further disproven as all seven sites are relatively far away from any continental
299 inputs, with the closest site (Site 7) located ~200 km away from the nearest coast.

300 Previous studies have determined that *Sargassum* spp. metal concentrations vary on
301 both spatial and temporal scales (Soerensen et al., 2014; Rodríguez-Martínez et al., 2020).
302 Therefore, the heterogeneity in metal(loid) concentrations observed between this study's
303 seven sites' might be linked to the origin of the *Sargassum* spp. and their journey along
304 different oceanic currents and/or to regions with varied proximities to coastal contaminated
305 areas. In 2018, the year of this study, high densities of *Sargassum* spp. were observed in the

306 Great Atlantic *Sargassum* belt (Wang et al., 2019). Following the modeling study of Wang et
307 al. (2019), *Sargassum* spp. located in the central Atlantic, like the ones in our study, likely
308 developed locally rather than from seed populations in the Sargasso Sea as proposed by
309 Fernández et al. (2017). It is possible that some of the *Sargassum* spp. may have come from
310 West Africa and bloomed in the central Atlantic, validating the role of the North Equatorial
311 Recirculation Region as a potential source region as proposed by Frank et al. (2016). This
312 hypothesis would suggest the non-influence of previous coastal contamination on the
313 *Sargassum* spp. in our study and would imply relatively low metal(loid) contamination. From
314 their sampling location, *Sargassum* spp. rafts probably traveled along the Loop Current and
315 Gulf Stream to finally enter the North Atlantic Ocean. Some *Sargassum* spp. might have been
316 transported directly into the North Atlantic following the Antilles Current (Wang et al., 2019)
317 while other rafts may have entered the Caribbean Sea (Putman et al., 2019). The Equatorial
318 Atlantic's ocean circulation dynamics play a central role in the transport of *Sargassum* spp.
319 into the Caribbean Sea. Once there, trade-winds are responsible for their beaching by
320 transporting the superficial waters towards the shore, therefore pushing *Sargassum* spp. rafts
321 towards the coast.

322 Despite the significant differences between site-specific *Sargassum* spp. elemental
323 concentrations, this present study concentrations can still be considered low. As such, these
324 sites are relatively homogenous compared to the large variability observed in the literature
325 review from coastal benthic *Sargassum* spp. (Table S3 and Figure 3). Indeed, most of the
326 coastal benthic *Sargassum* spp. metal concentrations (Cd, Cu, Fe, Ni, and Hg) are
327 significantly higher than values measured in our study. We have to take into consideration
328 that most of the sampling sites from the reviewed coastal benthic *Sargassum* spp. studies are
329 also from highly industrialized coastal areas. As presented above, external inputs that supply
330 trace metal(loid)s to oceanic surface waters, such as river runoff, have a strong effect on
331 coastal waters and therefore on the coastal *Sargassum* spp. metal contaminations.

332 The comparison between this study's open ocean planktonic *Sargassum* spp. element
333 concentrations to previously published coastal planktonic *Sargassum* spp. metal
334 concentrations show contrasting results (Table 1). Most *Sargassum* spp. metal concentrations
335 (Cd, Cu, Fe, Zn, and Hg) from the coast of Ghana's (Addico and Atta deGraft-Johnson, 2006)
336 present levels significantly higher (up to a factor of one hundred) than values from our study;
337 arsenic concentration are however significantly lower, by a factor of five. According to
338 Addico and Atta deGraft-Johnson (2006), their *Sargassum* spp. samples come from areas
339 associated with intensive mining and industrial activities that might explain the observed high

340 element concentrations. Compared to the results generated in this present study, metal and
341 metalloid concentrations from the Dominican Republic region (Fernández et al., 2017) are in
342 a similar range for Cd, Cu, Mn, Ni, V, and Zn; significantly higher by a factor of 20, 50, and
343 100 for Cr, Hg, and Al, respectively; and significantly lower by a factor of five for As. The
344 origin of the *Sargassum* spp. samples from the Dominican Republic seem to be more closely
345 linked to sources in the southern latitudes. There they might have proliferated near the mouth
346 of large rivers exposed to coastal contaminations that could explain their high metal and
347 metalloid concentrations (Fernández et al., 2017). The most recent study from Rodríguez-
348 Martínez et al. (2020) present *Sargassum* spp. element concentrations from the Mexican
349 Caribbean coast. Most of their metal and metalloid concentrations are in the same range as
350 presented in our study, even for As. The comparison to these recent studies leads one to
351 conclude that our new assessment of pelagic open ocean *Sargassum* spp. element
352 concentrations are on average in the same range or below already published planktonic
353 *Sargassum* spp. values from non-beached *Sargassum* spp. This validates the hypothesis that
354 pelagic *Sargassum* spp. from this study bloomed in the open ocean and have not yet
355 encountered potential contaminated coastal areas.

356 Regarding As, it is a metalloid naturally present in the ocean, in concentrations
357 between 15 and 25 nM for hydrogen arsenate (Millero, 2006). Various studies have concluded
358 that brown macro algae rapidly and greatly accumulate dissolved As (e.g. Penrose, 1974; Neff
359 et al., 1997, Devault et al., 2020). Both this present study with samples from the open ocean
360 and studies on the Mexican coast yield similar levels of As concentration, both higher than the
361 ones from the more contaminated coastal areas of Ghana (Addico and Atta deGraft-Johnson,
362 2006) and Dominican Republic (Fernández et al., 2017). In contaminated areas, heavy metals
363 are highly abundant and could consequently saturate fixation sites of *Sargassum* spp. This
364 competition for binding sites could reduce the number of available sites for As and would
365 explain the low As concentration recorded in *Sargassum* spp. from contaminated areas
366 compared to the open ocean.

367 A recent study tackling the ecotoxicology effect of nano-plastics on marine organisms
368 analyzed metal contamination in North Atlantic gyre micro-plastics (Baudrimont et al., 2020).
369 Metal contamination in micro-plastics (26.155 $\mu\text{g}\cdot\text{g}^{-1}$ for Fe; 3.364 $\mu\text{g}\cdot\text{g}^{-1}$ for Cu; 16.633 $\mu\text{g}\cdot\text{g}^{-1}$
370 $\mu\text{g}\cdot\text{g}^{-1}$ for Zn; 2.051 $\mu\text{g}\cdot\text{g}^{-1}$ for Ni; and 0.552 $\mu\text{g}\cdot\text{g}^{-1}$ for Cd) is within the same range yielded by
371 samples presented in this study's *Sargassum* spp. samples. This could indicate that *Sargassum*
372 spp. can adsorb metals in a manner similar to micro-plastics, or that the metal contamination
373 recorded in *Sargassum* spp. can be due to the presence of nano-plastics attached to them. A

374 more thorough study exploring these mechanisms would be interesting and necessary to
375 validate or refute these hypotheses.

376 Element concentrations in *Sargassum* spp. presented in this study are well below trace
377 elements limits of the French norm for the enrichment of organic soil product (NFU 44-051-
378 ISSN 0335-3931) for Cd, Cr, Hg, Ni, Pb, Se, Cu, and Zn. Only the concentration of As in
379 *Sargassum* spp. is above the acceptable value of 18 $\mu\text{g}\cdot\text{g}^{-1}$. This is not surprising as various
380 studies have concluded that brown macro algae rapidly and greatly accumulate dissolved As
381 (e.g. Penrose, 1974; Neff et al., 1997). However, our *Sargassum* spp. samples yielded a mean
382 As concentration (145.62 $\mu\text{g}\cdot\text{g}^{-1}$) that falls well within the worldwide range of 0.1 - 382 $\mu\text{g}\cdot\text{g}^{-1}$
383 for marine algal samples (Neff et al., 1997). That being said, the level of As enrichment might
384 become a concern for the use of *Sargassum* spp. in the industry as it may pose potential health
385 risks. The mean As concentration represents total As in *Sargassum* spp., however, it is already
386 known that only certain forms of As are toxic (e.g. Neff, 1997). Further study would be
387 needed to explore differences in the concentrations of the various forms of As that can be
388 found in *Sargassum* spp. in order to determine the true level of toxicity.

389 Open ocean *Sargassum* spp., such as the ones analyzed in this study, can therefore be
390 considered as relatively pristine as they did not yield high metal concentrations, aside from
391 the metalloid As. Concentrations from this study can be used as a first assessment for future
392 *Sargassum* spp. contamination studies that endeavor to tackle the issue of estimating coastal
393 contamination. For example, the element concentrations found in the Caribbean beaching
394 *Sargassum* spp. can be studied in the context of our open ocean *Sargassum* spp. samples.
395 While we acknowledge the fact that seven sites are not enough to have a comprehensive
396 understanding of all open ocean *Sargassum* spp. metal and metalloid concentrations, this
397 study still gives a meaningful first assessment. We also recognize the limitation of using a
398 mixture of two or three macroalgae species which can accumulate metal(oids) differently as
399 already pointed out by Milledge et al., (2000). Follow up studies with improved spatial and
400 temporal coverage as well as sample size will be necessary to obtain more detailed
401 calculations as previously developed for algae (Garca-Seoane et al., 2018) and bivalves (Lu et
402 al., 2019), to determine baseline open ocean metal(loid) loads.

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405 **V. Conclusions**

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407 This study revealed that Atlantic open ocean *Sargassum* spp. do not transport
408 significant metal(loid) loads and could be considered as pristine before reaching the
409 Caribbean Sea, with As as a potential exception that needs to be further explored. These
410 results are reassuring since free floating *Sargassum* spp. provide a habitat for many
411 organisms, including commercially relevant fish that could have been impacted by metal(loid)
412 enrichment. Although the seven sites present statistically significant site-specific differences,
413 the spatial sampling is not sufficient to determine the cause of this variability or to set
414 baseline metal(loid) concentrations representative of all open ocean *Sargassum* spp.
415 populations. Since *Sargassum* spp. present in the western Equatorial Atlantic have a high
416 probability of entering the Caribbean Sea within a year's time (Putman et al., 2019) our
417 studied *Sargassum* spp. rafts might have been transported to the Caribbean region, resulting in
418 massive beaching events. Despite the massive arrival of *Sargassum* spp., the upside is that it
419 is unlikely that these algae would transport any preexisting metal(loids) contamination. These
420 elemental concentrations recorded in the *Sargassum* spp. samples presented in this study can
421 be considered reference values for future work that focuses on *Sargassum* spp. from coastal
422 areas. Additionally, replication studies might be of high interest to further validate our first
423 assessment of open ocean element concentrations and sample sites in closer proximity to the
424 coast may elucidate a potential element enrichment in the *Sargassum* spp. sampled in sites
425 closer to the coast.

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428

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434 spp. samples.

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438 **Ethical Approval**

439 This manuscript is an original work and has not been previously published somewhere else
440 nor submitted to more than one journal for simultaneous consideration.

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Consent to participate

Not applicable

Consent for publication

Not applicable

Author contribution

EPD carried out sample preparation, analyses and data acquisition; executed the analytical research and interpretation, and served as primary author. PYG carried out statistical analyses. OAC accomplished *Sargassum* species identification. MB obtained the *Sargassum* samples. PYP and MB proofread the manuscript. All authors contributed to the article and approved the submitted manuscript version.

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Availability of data and materials

All the data discussed in this study are the one presented in this manuscript's tables.

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